

1 **Greenhouse Gas Accounting Procedures in Low Carbon Fuel Policies**
2 **Lead to Undervalued Benefits of Miscanthus-based Sustainable**
3 **Aviation Fuel**

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21 **Keywords**

22 sustainable aviation fuel (SAF); ethanol-to-jet (ETJ); alcohol-to-jet (ATJ); life cycle assessment
23 (LCA); techno-economic analysis (TEA); cellulosic biofuel; measurement, monitoring, reporting,
24 and verification (MMRV)

25 **Abstract**

26 Low carbon fuel policies such as the U.S. Renewable Fuel Standard (RFS), Canada Clean Fuel
27 Regulations (CFR), and California Low Carbon Fuel Standard (LCFS) are intended to reduce the
28 greenhouse gas (GHG) emissions from transportation. Cellulosic feedstocks, optimized
29 biorefineries, and favorable farming locations can significantly reduce biofuel carbon intensity (CI).
30 Despite the emergence of field-to-fuel GHG monitoring technologies that could verify such
31 benefits, programmatic constraints in CI accounting procedures may limit fuel producers' ability
32 to capitalize on these opportunities. To elucidate the implications of this challenge, this work
33 examines a miscanthus-to-sustainable aviation fuel (SAF) pathway (i) to demonstrate how
34 program provisions drive estimates of biofuel CIs and (ii) to explore potential CI and financial
35 benefits of spatially explicit life cycle assessment (LCA). In comparing policy-based vs. spatially
36 explicit CI scores (estimated via DayCent and BioSTEAM) for SAF production from miscanthus
37 via alcohol-to-jet (ATJ), programmatic CI accounting requirements underestimated GHG benefits
38 in 60-99% of simulated scenarios. These underestimates result in policy-induced SAF price
39 differentials of -1.19 [(-)3.46 to (-)0.23], -0.07 [(-)1.06 to (+)0.37], and -0.48 [(-)2.46 to (+)0.16] \$·L⁻¹
40 for the RFS, CFR, and LCFS, respectively. Ultimately, this work demonstrates the importance
41 of LCA methodological specifications in low carbon fuel policies.

42 Introduction

43 Low carbon fuel policies such as the U.S. Renewable Fuel Standard (RFS), Canada Clean Fuel
44 Regulations (CFR), and California Low Carbon Fuel Standard (LCFS) are intended to reduce the
45 greenhouse gas emissions associated with transportation. The RFS is unique among these three
46 programs because of its pathway-based fuel carbon intensity (CI) reduction threshold
47 categorization structure (sometimes referred to as a “technology mandate”). If fuel producers
48 verify they utilized a certain feedstock and production process to produce a certain fuel, the fuel
49 is assigned to the corresponding fuel category (referred to with a “D code”) and effectively
50 assigned a predetermined CI (e.g., a D code of 3 or 7 signifies that the fuel was produced from a
51 cellulosic feedstock and achieves a 60% CI reduction compared to a petroleum baseline fuel¹).
52 The CFR and LCFS are “performance-based,” and allow fuel producers to calculate unique CIs
53 for their own production pathway. Relative to a technology mandate, a fuel’s performance in these
54 programs has the potential to be more closely (but still not fully) tied to the actual CI reduction
55 achieved by the fuel. The LCFS is credited with reducing the CI of California’s transportation fuel
56 pool by approximately 15%,² and U.S. bioethanol production has increased roughly 8-fold over
57 the past two decades to meet RFS requirements.³ However, cellulosic biofuel production targets
58 set by the RFS have never been met, thus preventing the realization of the additional benefits of
59 these fuels compared to conventional fuels (e.g., improved biodiversity in agricultural
60 landscapes,⁴ sequestration of atmospheric carbon in soils,⁵ reduced human health externalities⁶).
61 Critiques of the RFS suggest that the technology mandate does not properly incentivize the
62 production of fuels that achieve lower CIs.⁷ However, even performance-based programs may
63 not necessarily account for production decisions that affect a fuel’s CI at all stages of biofuel
64 production.

65 The CFR is performance-based, yet requires the use of a single value for the CI of crop
66 production for many feedstocks, regardless of the location where production took place or the
67 methods employed (and despite data that underly the CFR illustrating that the CI of crop
68 production varies widely across Canada for many crops^{8–10}). The LCFS also does not allow fuel
69 producers complete flexibility to determine a unique field-level feedstock CI by accounting for
70 field-specific biogeochemistry. However, yield and soil carbon sequestration by miscanthus, a
71 perennial grass energy crop that can produce significant biomass with few chemical inputs, have
72 been estimated to vary from approximately 7 to 22 megagrams per hectare and -1.0 to 1.5
73 megagrams of carbon per hectare-year, respectively, across the rainfed U.S.⁵

74 Life cycle assessment (LCA) is a quantitative methodology regularly used to evaluate the
75 environmental impacts of biofuel production, including as part of low carbon fuel policies and tax
76 credits. In short, LCA entails combining quantities of material and energy flows in a product or
77 process system (e.g., fertilizers, electricity, and biofuels in the case of biorefineries) with
78 information on their likely environmental impacts to determine the overall environmental impacts
79 of the product or process (e.g., according to ISO Standards^{11,12}). Depending on the goal and
80 scope of a LCA and data availability, the assessment may utilize emissions data from different
81 scales (e.g., national averages vs. regional vs. state-level). There is, however, a recognition that
82 assessments conducted using low-granularity (i.e., not site-specific) data may not properly
83 elucidate differences in environmental impacts that inevitably occur across individual production
84 sites.¹³ Thus, there has been movement toward using more site-specific (e.g., “spatially explicit”,
85 “location-specific”, “field-level”) data in LCAs, particularly for analyses of bioenergy crops,
86 biorefining technologies, and bioproducts. Simultaneously, improved technologies to facilitate
87 field-level measurements of important parameters are being developed to support these analyses
88 (i.e., improved measurement, monitoring, reporting, and verification [MMRV]). For example, a
89 proposed combination of field-specific modeling and sensing and big data analytics has the
90 potential to more accurately quantify agricultural greenhouse gas (GHG) outcomes.¹⁴ Given the
91 continued interest in emerging fuel production pathways (e.g., miscanthus to sustainable aviation
92 fuel [SAF] as a means to decarbonize medium- and long-haul aviation travel¹⁵⁻¹⁸), there exists a
93 need to evaluate the implications of such field-specific data for biofuel CI quantification and, in
94 particular, a comparison to CIs calculated based on existing policy guidelines. Understanding the
95 flexibility of policy-specified CI quantifications such as those utilized by the Clean Fuel Production
96 Credit (a tax credit also known as the 45Z credit because it is established in 26 U.S. Code § 45Z)
97 is critical to capitalize on the benefits of biofuels and field-specific monitoring.

98 The overarching goal of this work is to explore how programmatic constraints may affect
99 estimates of biofuel CIs for emerging feedstock-to-fuel pathways and to highlight the potential
100 benefits of implementing field-specific LCAs of biofuels. To illustrate the importance of sound
101 policy provisions and highlight the potential utility of field-to-fuel GHG monitoring technology, the
102 carbon intensity of SAF produced from miscanthus via the alcohol-to-jet (ATJ) pathway was
103 calculated according to specifications from three low carbon fuel policies (the RFS, CFR, and
104 LCFS) and the 45Z tax credit and compared against spatially explicit modeling results tailored to
105 the feedstock-to-fuel pathway. The CI for the spatially explicit analysis was compared to policy-
106 specific CIs to illustrate how policy guidelines and current flexibility affect fuel characterization for
107 this potential pathway. In addition to more accurately representing true environmental outcomes

108 of biofuel production, policies that enable field-specific CI calculations could better incentivize
109 compliance and impactful carbon mitigation measures by producers.

110

111 **Methods**

112 **Feedstock-to-Fuel Pathway**

113 Miscanthus is a high-yielding, perennial grass with significant potential to produce biomass for
114 conversion to fuels because it requires little inputs of fertilizers and pesticides.^{19–22} There is
115 particular potential for miscanthus production (for eventual conversion to biofuel) east of the 100th
116 meridian in the United States because this area can rely on rainfed agriculture.⁵ Miscanthus is
117 one of several cellulosic feedstocks that can be used to produce ethanol, though this process
118 remains expensive.^{23,24} Cellulosic ethanol can be further upgraded to SAF (potentially in the same
119 biorefinery) to produce a more cost-competitive fuel product.^{25,26} Though SAF production is in
120 early stages in the U.S.,²⁷ this ATJ pathway is one of 11 SAF production pathways approved by
121 ASTM International.²⁸

122 **Policy Provisions**

123 *Policies and Specifications for LCA*

124 The three low carbon fuel policies evaluated in this study share similar goals: to require the
125 blending of fuels with lower greenhouse gas emissions (RFS;²⁹ see § 80.2), to prevent or reduce
126 air pollution from fossil fuels (CFR;³⁰ see the Registration section), and to reduce the carbon
127 intensity of transportation fuels (LCFS;³¹ see § 95480). With regard to the 45Z tax credit, one of
128 the primary goals of the Inflation Reduction Act of 2022 (P.L. 117-169) was to address climate
129 change by incentivizing the growth of a clean energy economy: the 45Z tax credit was included
130 by Congress to further incentivize the production of biofuels with lower carbon scores or GHG
131 reductions.³² The three programs and the tax credit all specify methods to determine the CI of
132 biofuels produced for program compliance, which includes quantifying the CI of producing the
133 biomass feedstock used to produce the fuel (detailed in RFS § 80.1426, in CFR § 75-76, in LCFS
134 § 95488.3, and in 45Z § (b)(1)(B)(iii)). All three programs and the tax credit utilize an LCA
135 approach for CI quantification, with slight variations in scope; therefore, the policy, and the
136 methods it specifies, have some influence in determining the CI that is calculated for a given
137 biofuel from a given field. The methods specified by each policy were reviewed to highlight which
138 policy choices might affect biofuel CI estimates. Attributes of policies considered include the use

139 of site-specific data for each life cycle stage and any material and energy flows or other
140 considerations that influence CI calculations.

141 Policy-Based Miscanthus-to-SAF CI Calculations

142 Policy-based SAF CIs were calculated according to methods specified by the four policies. In the
143 case of the RFS and 45Z credit, which generally do not allow for producer-specific CIs to be used,
144 predetermined CIs developed by program administrators were used as the policy-based CIs. For
145 the CFR and LCFS, policy-based CIs were calculated according to the methods specified by the
146 policies (to the greatest extent possible given current data and model availability and the use of
147 simulated biorefinery operation data). Policy-based CIs for all four policies are detailed further in
148 SI Section S1.1.

149 **Farm-to-Fuel Life Cycle Assessment**

150 Spatial Variability in Feedstock CI

151 CIs of miscanthus production were obtained for 36 U.S. states from Fan *et al.* (2024).³³ Fan et al.
152 estimated the CI of miscanthus production across the rainfed U.S. using DayCent to account for
153 spatially-varying growing conditions and weather and region-specific historical cultivation
154 practices (detailed further in the Supporting Information for ³³). The DayCent model used
155 parameters including precipitation, fertilizer application, land cover, and land management to
156 calculate miscanthus yield, changes in soil organic carbon (SOC), and in-field greenhouse gas
157 emissions associated with chemical inputs. The miscanthus CI was calculated as the sum of SOC
158 change and in-field and upstream (scope 3) emissions associated with chemical inputs (nitrogen,
159 phosphorus, and potassium fertilizers; herbicides; and fuel) divided by miscanthus yield.

160 Notably, Fan *et al.* did not vary the rate of fertilizer and herbicide application, though these
161 parameters are likely to vary across the U.S. and from year-to-year. Additionally, there are
162 methodological differences necessary to account for SOC compared to these non-SOC emissions
163 (i.e., while quantities of fertilizer inputs and environmental impacts associated with their
164 production may be measured fairly directly, long-term miscanthus-induced SOC storage must
165 currently be estimated using biogeochemical models, as in ³³). As estimated by Fan et al., non-
166 SOC emissions are relatively minor compared to the magnitude of carbon sequestered by
167 miscanthus growth (15.0% [8.8-37.6%] of SOC sequestration). For transparency, the contribution
168 of SOC and non-SOC emissions to the total SAF CI are reported separately in this study. These
169 feedstock CIs are meant to represent the potential range of CIs that may be observed with robust

170 field-level MMRV. The CIs of miscanthus production were incorporated into the BioSTEAM
171 Location Specific Evaluation module (BLocS).^{34,35} The median miscanthus CI was -139 kg
172 CO₂e-dry metric ton⁻¹, indicating a net removal of carbon from the atmosphere and storage as
173 SOC.

174 Transportation and Distribution CI

175 The CI associated with biomass and biofuel transportation and distribution was estimated by
176 multiplying the mass of biomass or biofuel and transport distance by transportation emission
177 factors for a diesel truck (expressed in kg CO₂e·kg·km⁻¹) from ecoinvent 3.8 and the TRACI 2.0
178 impact assessment methodology. Because there are only two operational SAF refineries and zero
179 operational cellulosic ethanol refineries in the U.S.,³⁶ a range of biomass and biofuel transport
180 distances (see SI **Table S1**) were evaluated to characterize the potential consequences of this
181 parameter on the overall CI of SAF production. The median biomass and biofuel transport
182 distances were 100 and 600 kilometers, respectively, which accounts for two trips (i.e., one full
183 transport trip and one empty return trip). The CI of both transportation and distribution stages was
184 added to the CI of the biomass conversion stage calculated using BioSTEAM and BLocS.

185 Biorefinery Design, Simulation, and LCA

186 Biorefinery design, simulation, and LCA were performed using BioSTEAM version 2.44.3.^{37,38}
187 BioSTEAM compiles a life cycle inventory for the biomass conversion stage based on process
188 simulation, then combines this with user-defined emission factors to perform impact assessment.
189 This study leveraged BLocS for location-specific LCA inputs³⁴ (electricity and natural gas emission
190 factors based on balancing region and state, respectively). Life cycle inventory data for chemical
191 inputs to the biomass conversion stage that do not contribute significantly to the total fuel CI and
192 are unlikely to vary significantly with location (e.g., sodium hydroxide) were obtained from
193 ecoinvent 3.8. The influence of the full range of miscanthus CIs across locations on the total fuel
194 CI was evaluated via LCA.

195 The SAF model in BioSTEAM simulates the production SAF from cellulosic biomass (i.e.,
196 miscanthus) via the ATJ pathway, consistent with Wei et al.³⁹ The SAF production process
197 includes two steps: ethanol production and ethanol upgrading via dehydration, oligomerization,
198 and hydrogenation. The biorefinery uses 0.821 million metric tons of miscanthus per year and
199 produces several hydrocarbons: SAF as the main product and gasoline and diesel as co-products.
200 Excess electricity produced from burning waste biomass is sold to the grid.

201 Biofuel Combustion

202 The carbon intensity of biofuel combustion was excluded from consideration as part of farm-to-
203 fuel LCA. The three low carbon fuel policies evaluated in this study consider carbon dioxide
204 emissions from biofuel combustion to be biogenic and offset by carbon uptake during biomass
205 production and thus exclude them from consideration, though the RFS and CFR consider other
206 greenhouse gases associated with combustion. The CI of fuel combustion would not vary based
207 on production location or policy participation.

208 Uncertainty

209 While uncertainty in biofuel LCA was considered when developing the three low carbon fuel
210 policies, none of the programs allow participants to report a biofuel CI with an uncertainty range.
211 Under the LCFS, fuel producers are encouraged to report a slightly higher CI than they calculate
212 to account for uncertainty in biorefinery performance (because they are penalized if the actual
213 fuel CI is determined to be higher than what they report to the program). Uncertainty was
214 considered for all life cycle stages (except fuel combustion; detailed in SI **Tables S1-2**) (1) to
215 provide a more accurate estimate of the total fuel CI and (2) to illustrate the potential range of fuel
216 CIs that are currently unaccounted for under program provisions. Latin hypercube sampling was
217 used to generate samples for Monte Carlo simulation. Ten-thousand samples were generated for
218 the analysis.

219 **Influence of Policy Specifications on Fuel Characterization**

220 Techno-Economic Analysis (TEA)

221 BioSTEAM and BLocS were also used to estimate the minimum SAF selling price via techno-
222 economic analysis (TEA). BLocS contains state-specific economic parameters (e.g., miscanthus
223 prices, tax rates, capital cost factors) that are incorporated into TEA by BioSTEAM. Miscanthus
224 prices in BLocS were obtained for 31 U.S. states from Lee et al. (2023).⁵ The influence of the full
225 range of miscanthus prices across locations on the minimum SAF selling price was evaluated via
226 TEA.

227 Policy-Induced Fuel CI and Price Differentials

228 To illustrate the influence of low carbon fuel program provisions on biofuel CI determination, the
229 policy-induced CI differential was calculated by subtracting the CI calculated according to program
230 specifications from the CI calculated via spatially explicit LCA (**Equation 1**). Therefore, a negative
231 CI differential indicates that the policy overestimates the fuel CI, and vice versa.

232

$$233 \quad differential_{CI} [g CO_2e \cdot MJ^{-1}] = CI_{LCA} [g CO_2e \cdot MJ^{-1}] - CI_{policy} [g CO_2e \cdot MJ^{-1}] \quad \text{Equation 1}$$

234 To further illustrate the effect of program provisions on fuel characterization and valuation,
235 the policy-induced price differential was calculated by multiplying the policy-induced CI differential
236 by the policy-specific credit selling price and adjusting units using the lower heating value (in
237 MJ·kg⁻¹) and density (in kg·L⁻¹) of SAF (**Equation 2**). A negative price differential indicates that
238 policy provisions contribute to an undervaluation of a fuel, and vice versa.

239

$$240 \quad differential_{price} [\$ \cdot L^{-1}] = differential_{CI} [g CO_2e \cdot MJ^{-1}] \times credit\ price [\$ \cdot g CO_2e^{-1}] \times \\ 241 \quad 44.15 [MJ \cdot kg^{-1}] \times 0.7507 [kg \cdot L^{-1}] \quad \text{Equation 2}$$

242

243 RFS credit selling prices (i.e., renewable identification number [RIN] selling prices) were
244 obtained from the EPA.⁴⁰ LCFS credit prices were obtained from the California Air Resources
245 Board.⁴¹ Because the CFR was only established in 2022, the credit market is nascent and no
246 price information is available. LCFS credit prices were used to estimate the CFR-induced price
247 differential due to the relative similarity of the programs. The ranges of credit selling prices from
248 the years 2014 through 2023 (adjusted to 2023 dollars, detailed in SI **Table S3**) were used to
249 calculate the price differentials.

250

251 **Results & Discussion**

252 **Comparison of Low Carbon Fuel Program Provisions and LCA Scopes**

253 *U.S. Renewable Fuel Standard*

254 The U.S. Renewable Fuel Standard (RFS) was originally created in 2005 by the Environmental
255 Protection Act and was amended and expanded in 2007 by the Energy Independence and
256 Security Act. The RFS mandates that certain amounts of biofuels are blended with conventional
257 fossil fuels used for transportation in the U.S., with the obligation to use biofuels placed on refiners
258 producing gasoline or diesel fuel. These so-called obligated parties are required to submit proof
259 of biofuel blending as renewable identification numbers (RINs), which are associated with batches
260 of biofuels when they are produced and separated upon blending. RINs may also be traded

261 between obligated parties with surpluses and deficits, and this trading creates a RIN market that
262 determines the selling price. The extra profits from sales of RINs are meant to incentivize the
263 production of biofuels, with the overall purpose of the RFS program being to reduce the
264 greenhouse gas emissions associated with transportation fuels. In actuality, the RIN market has
265 been extremely volatile and production goals for biofuels from cellulosic feedstocks have never
266 met RFS mandates. The majority of biofuel produced for RFS compliance has been ethanol from
267 corn grain, which has dubious environmental benefits.⁴²

268 The RFS is unique among the three programs considered here in that, for a given
269 feedstock and fuel production pathway, it categorizes fuels based on their achievement of CI
270 reduction thresholds rather than their actual, producer-specific CI (**Table 1**). For example, all
271 ethanol produced from cellulosic feedstocks is considered to achieve a 60 percent reduction in CI
272 compared to gasoline. These thresholds are based on LCAs performed by the EPA.⁴³ The EPA
273 states that its feedstock analyses are “country-neutral” and apply regardless of the growing
274 practices used at individual farms because the environmental impacts of feedstock production
275 “are likely to be the same regardless of which farm grows the specific feedstock used for biofuel
276 production.”⁴⁴ However, critiques of the RFS cite this threshold-based system as not properly
277 incentivizing the production of biofuels that achieve a greater reduction in CI (i.e., because two
278 cellulosic biofuels that achieve a 60% reduction and a 80% reduction, respectively, are considered
279 identical in terms of program compliance).

280 Canada Clean Fuel Regulations

281 The Canada Clean Fuel Regulations (CFR) were established in 2022 under the authority of the
282 Canadian Environmental Protection Act of 1999. Similar to the RFS, the CFR requires producers
283 of gasoline and diesel, referred to as primary suppliers, to lower the overall CI of fuels they
284 produce by utilizing biofuels. Biofuel producers also generate compliance credits that can be
285 traded to create a market to further incentivize biofuel production. The CFR was implemented to
286 prevent or reduce air pollution from the combustion of liquid fossil fuels.³⁰

287 The CFR differs from the RFS in that producers have the option to calculate a CI for a
288 unique fuel production pathway via LCA. Calculations to determine the CI of fuels produced for
289 CFR compliance must be performed in the Government of Canada’s Fuel LCA Model,⁴⁵ which is
290 based in the openLCA software.⁴⁶ However, while biomass conversion processes may readily be
291 customized based on the producer’s unique methods, the CFR requires the use of a single CI for
292 biomass feedstocks including corn, sorghum, sugarcane, and corn stover, among others (**Table**

293 1).⁴⁷ While it is possible for a new feedstock CI to be approved for use in calculations, the Fuel
294 LCA Model guidance currently states: “Feedstock production in a subregion of a region included
295 in the [Fuel LCA Model] Data Library is not an eligible criterion for submitting a new pathway
296 application.”⁴⁷ Thus, the CIs for these crops available to use in the Fuel LCA Model are weighted
297 averages of several provincial-level CIs for each crop.⁴⁸ Miscanthus is not a feedstock currently
298 included in the Fuel LCA Model, though fuel produced from this crop would be eligible for a new
299 pathway application including the unique crop CI. However, given the treatment of other crop CIs
300 by the CFR, a CI for miscanthus production included in the Fuel LCA Model would likely represent
301 the average CI of production, across the U.S., for example.

302 California Low Carbon Fuel Standard

303 The California Low Carbon Fuel Standard (LCFS) was implemented in 2011 pursuant to the
304 California Global Warming Solutions Act of 2006. The LCFS is a state, rather than national,
305 program meant to reduce the carbon intensity of transportation fuels used in California.³¹ As with
306 the RFS and CFR, so-called fuel reporting entities are required to demonstrate that they produced
307 and blended biofuels by retiring compliance credits. However, the LCFS has often been cited as
308 better incentivizing low-CI fuels (e.g., fuels produced from cellulosic biomass) compared to the
309 RFS because it requires the calculation of a unique fuel CI rather than relying on reduction
310 threshold categories (and for this reason the developers of the CFR referenced the LCFS).

311 The LCFS further differs from both the RFS and CFR by allowing fuel producers to input
312 site-specific (or “pathway-specific”) values for some LCA parameters related to the feedstock
313 production stage (e.g., fertilizers; **Table 1**). Calculations to determine the CI of fuels produced for
314 LCFS compliance must be performed in simplified calculators or using CA-GREET.⁴⁹ However,
315 these models rely on non-site-specific assumptions for other key factors related to feedstock
316 production (e.g., emissions of greenhouse gases such as N₂O from fertilized cropland). The LCFS
317 may allow for further adjustment of feedstock production CIs to include site-specific
318 biogeochemistry (e.g., Table D.2 [row 2.4] in the Tier 1 Simplified CI Calculator for Biodiesel and
319 Renewable Diesel Instruction Manual⁵⁰ states: “if user-defined is selected [for the GHG emission
320 factor of soybean farming], consult CARB staff to develop emission factors...”), but this option is
321 not fully explained, guaranteed, or readily available to fuel producers.

322 26 U.S. Code § 45Z Clean Fuel Production Credit

323 The 45Z tax credit was established in 2022 by the Inflation Reduction Act and will be offered to
324 U.S. taxpayers as of January 1, 2025.⁵¹ The tax credit encourages the production of low emission

325 transportation fuels by offsetting qualifying taxpayers' federal tax liability. Taxpayers interested in
 326 claiming this credit must prove to the U.S. Internal Revenue Service that they produced or sold
 327 transportation fuels with lower greenhouse gas emissions than conventional fossil-based fuels. In
 328 the case of SAF, the credit is worth 0.09 $\text{\$.L}^{-1}$ or 0.46 $\text{\$.L}^{-1}$, depending on whether the fuel is
 329 produced according to certain wage and apprenticeship requirements, multiplied by a factor that
 330 compares the SAF CI to a baseline fuel CI (see ⁵² for additional details).

331 Though the credit has not yet been fully implemented, it appears structured to categorize
 332 fuels and assign them predetermined CIs (published by the U.S. Secretary of the Treasury) that
 333 are not producer-specific (**Table 1**). However, in the case of SAF, the CI must be calculated in
 334 accordance with CORSIA (Carbon Offsetting and Reduction Scheme for International Aviation)
 335 or a similar methodology. Default biofuel CIs calculated for CORSIA include biomass feedstock-
 336 induced SOC storage.⁵³ Thus, while it does not appear that the 45Z credit will allow taxpayers to
 337 calculate unique SAF CIs for participation, the program does consider farming inputs and crop
 338 field biogeochemistry.

339 **Table 1.** Summary of the scope of LCAs for low carbon fuel policies and the 45Z tax incentive.

	Biomass production	Biomass transportation & distribution	Biomass conversion	Biofuel transportation & distribution	Biofuel combustion
RFS	all life cycle stages are included, but program performance is technology- and threshold-based and use of a site-specific CI is generally not possible				
CFR	biomass CIs based on LCAs available in CFR data library LUC effects excluded (except tillage practices and changes in summer fallow area)	distance and mode adjustable	adjustable for some fuels only hydrogen and electricity input considered for SAF denaturant is excluded	distance and mode adjustable	biofuel combustion CIs based on LCAs available in CFR data library biogenic emissions excluded
LCFS	farming inputs generally adjustable (depends on feedstock and Tier 1/2 calculator/model) biogeochemistry not adjustable LUC included	distance and mode adjustable	fully adjustable	distance and mode adjustable	zero combustion emissions
45Z	all life cycle stages are included (as is biomass feedstock-induced soil organic carbon storage), but predetermined CIs are the default and the use of site-specific CIs appears unlikely to be possible				

340

341

342 **Biofuel CI and Relative Contributions of Production Stages**

343 Compared to the CI of conventional jet fuel (89.0 to 124.9 g CO₂e·MJ⁻¹), nearly all of these
344 estimates represent a CI reduction. The policy-specific SAF CIs (for the conversion of miscanthus
345 to SAF via ethanol) range from -51.3 (CFR, lowest CI of three scenarios) to 36.8 g CO₂e·MJ⁻¹
346 (RFS; **Table 2**). The maximum SAF CIs found in low carbon fuel documentation was 100.5 g
347 CO₂e·MJ⁻¹, which was associated with the most conservative temporary LCFS pathway CI. In
348 general, the policy-specific CIs tend to be higher than those estimated via spatially explicit LCA:
349 -100.7, -75.5, and -8.9 g CO₂e·MJ⁻¹ according to system expansion, hybrid allocation, and energy-
350 based allocation, respectively.

351 The SAF CI utilized for RFS compliance (36.8 g CO₂e·MJ⁻¹) represents a 60 percent
352 reduction in greenhouse gas emissions compared to conventional diesel fuel. This reduction
353 threshold is meant to be broadly representative of the various pathways for producing diesel and
354 related fuels from biomass (which result in different fuel CIs, according to EPA LCAs⁴³), and thus
355 cannot be specifically representative of a miscanthus-to-SAF pathway that has the potential to
356 result in a significantly lower CI. Two EPA LCAs⁴³ that were used to set the RFS compliance
357 threshold (though they represent a vegetable oil-to-SAF pathway) estimated an average CI of
358 32.9 g CO₂e·MJ⁻¹. Similarly to the RFS, the temporary SAF CIs (50.0 to 100.5 g CO₂e·MJ⁻¹; **Table**
359 **2**) available for LCFS participation (without a new approved pathway) are meant to be
360 representative of multiple technologies and present a very conservative estimate of emissions.⁵⁴
361 However, if a new pathway were approved for participation, the LCFS-based CI (-31.8 g CO₂e·MJ⁻¹)
362 could be closer to estimates following CFR methods and spatially explicit, energy-allocated
363 LCA.

364 Regardless of policy provisions or fuel production scenario, the feedstock production and
365 feedstock conversion stages are the main contributors to the total SAF CI (**Table 2**, **SI Figure**
366 **S1**). According to methods specified by the CFR, the feedstock production stage contributes -
367 10.4 g CO₂e·MJ⁻¹ to the total SAF CI. This result is comparable to the results of spatially explicit
368 LCA on a hybrid allocation basis, where this stage contributes -16.0 g CO₂e·MJ⁻¹ to the total -75.5
369 g CO₂e·MJ⁻¹ CI (median values). In all cases, the negative emissions during feedstock production
370 stem from modeled increases in soil organic carbon. The relative similarity of the CFR-based and
371 spatially explicit LCA CIs illustrates the benefits of accounting for key site-specific production
372 factors (e.g., farming inputs and biogeochemistry in the case of the miscanthus CI used for these
373 two analyses).

374 The feedstock conversion stage contributes -34.5 and -39.9 g CO₂e·MJ⁻¹ to total SAF CIs
375 according to CFR and LCFS methods, respectively. The key driver for negative emissions from
376 feedstock conversion is excess electricity production at the biorefinery. According to spatially
377 explicit LCA and hybrid allocation, this stage contributes a median of -62.0 g CO₂e·MJ⁻¹ (**Table**
378 **2**). One explanation for the difference in CIs is that the Fuel LCA Model used to calculate CIs for
379 CFR participation uses one model for the conversion of various feedstocks (including agricultural
380 residues, animal fats, and wood chips, among others) to SAF; users are only required to input
381 quantities of hydrogen used for production,⁴⁷ as opposed to the more rigorous process model and
382 life cycle inventory employed for this spatially explicit LCA (**Table 1**).

383 The other three life cycle stages (feedstock transportation and distribution, fuel
384 transportation and distribution, and fuel combustion) do not contribute significantly to the total
385 SAF CI (**Table 2**). Given the contribution of the miscanthus production stage to the overall SAF
386 CI, the fact that low carbon fuel policies generally limit the use of field-specific feedstock CIs
387 represents a critical challenge to capturing the anticipated benefits of this herbaceous energy crop.
388 Regardless of rationale for this limitation, the variation in feedstock production CIs at the field level
389 and the significance of this stage to the overall fuel CI highlights the importance of accounting for
390 this variation via spatially explicit (or even field-specific) LCA.

391

392 **Table 2.** Carbon intensity (CI) of SAF derived from miscanthus (via conversion to ethanol followed by ATJ)
 393 according to various policies and methodological choices. Total CI scores include miscanthus production,
 394 biomass conversion, feedstock transportation and distribution, fuel transportation and distribution, and fuel
 395 combustion.

Relevant Policy	Calculation and Pathway Description	Total CI [g CO ₂ e·MJ ⁻¹]	Miscanthus Production CI [g CO ₂ e·MJ ⁻¹]	Biomass Conversion CI [g CO ₂ e·MJ ⁻¹]
--	conventional jet fuel	+124.9 ^[a] +87.89 [LCFS 2024] +87.9 [CFR] +89.5 [45Z]	--	--
-- (this study)	spatially explicit LCA, system expansion; miscanthus feedstock	-100.7 [(-)319.5 to (+)9.0] ^b	-21.3 [(-)54.0 to (+)0.8] ^b	-82.6 [(-)284.5 to (+)11.6] ^b
	spatially explicit LCA, hybrid allocation; miscanthus feedstock	-75.5 [(-)239.5 to (+)6.8] ^b	-16.0 [(-)40.5 to (+)0.6] ^b	-62.0 [(-)213.3 to (+)8.7] ^b
	spatially explicit LCA, energy-based allocation; miscanthus feedstock	-8.9 [(-)45.7 to (+)21.1] ^b	-11.1 [(-)28.1 to (+)0.4] ^b	+1.9 [(-)23.8 to (+)19.4] ^b
RFS	60% reduction compared to conventional diesel average of the results of 2 EPA LCAs; vegetable oil feedstocks	+36.8 +32.9	N/A +17.7 ^c	N/A +14.2 ^c
CFR	Fuel LCA Model result; median U.S. miscanthus feedstock (would require new pathway application)	-42.9	-10.4	-34.5
	Fuel LCA Model result; 5 th percentile U.S. miscanthus feedstock (would require new pathway application)	-51.3	-18.7	-34.5
	Fuel LCA Model result; 95 th percentile U.S. miscanthus feedstock (would require new pathway application)	-36.6	-4.0	-34.5
LCFS	temporary CI; fats/oils/grease (FOG) residue feedstocks	+50	not specified	not specified
	temporary CI; plant oils (excluding palm oil) feedstocks	+70	not specified	not specified
	temporary CI; any other feedstocks	+100.5	not specified	not specified
	average of 26 approved producer pathways; waste oils, animal fats, and vegetable oils	+31.06	not specified	not specified
	REET Aviation Module result including default SOC change; miscanthus feedstock ^d	-31.8	+5.0	-39.9
45Z	CORSIA default CI for U.S. miscanthus ethanol-to-jet (ETJ) SAF	-23.6	-38.4	+12.3

396 ^a This value is calculated based on the percentage of long- and medium, and short-haul flights, and the percentage
 397 and carbon intensity of the energy source.^{55,56}

398 ^b Values shown represent median values followed by 5th and 95th percentiles (shown in brackets) of simulations across
 399 locations under uncertainty.

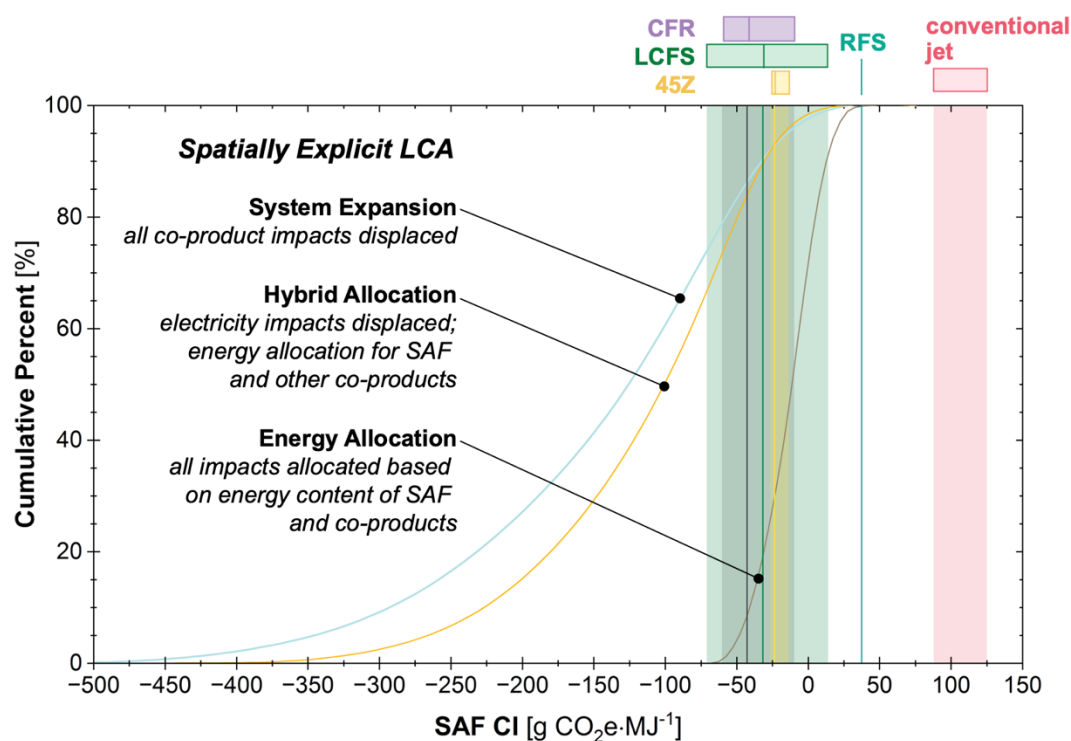
400 ^c These values are for vegetable oil feedstocks (not miscanthus), but are included here for comparison.⁴³

401 ^d This is representative of potential CA-REET development.

402

403 **The Potential for LCA to More Accurately Quantify Field-Specific GHG Outcomes**

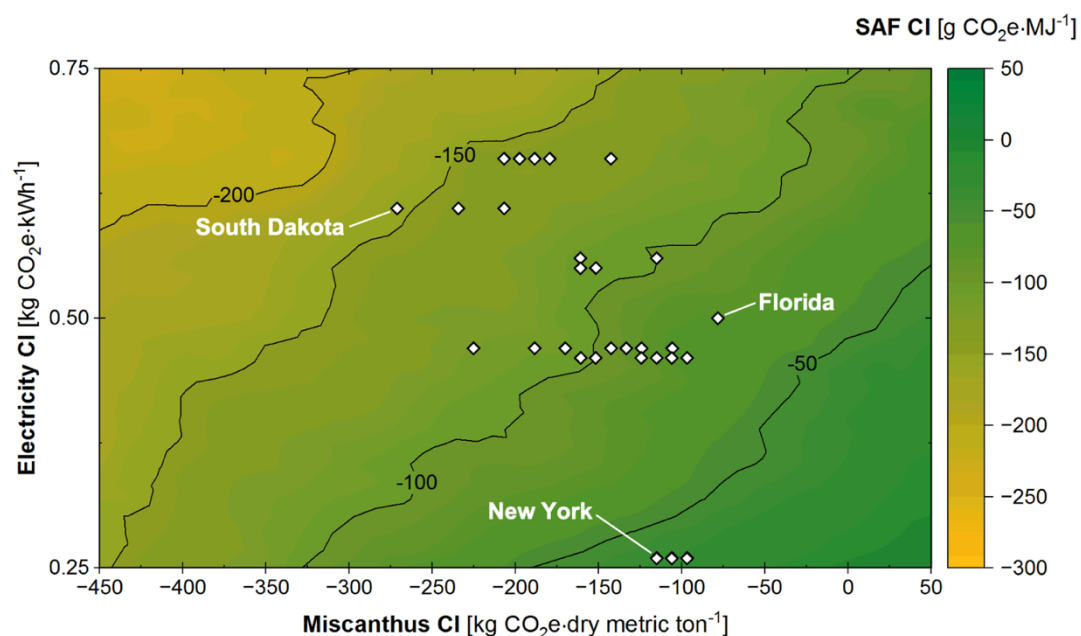
404 The CI of SAF production is subject to significant variation and uncertainty based on spatially
405 explicit LCA and tends to be overestimated by low carbon fuel policies (**Figure 1**). The SAF CI
406 calculated via spatially explicit LCA had a median value of $-100.7 \text{ g CO}_2\text{e}\cdot\text{MJ}^{-1}$ and 5th to 95th
407 percentiles of $[(-)319.5 \text{ to } (+)9.0]$ according to system expansion, $-75.5 \text{ g CO}_2\text{e}\cdot\text{MJ}^{-1}$ $[(-)239.5 \text{ to}$
408 $(+)6.8]$ according to hybrid allocation, and $-8.9 \text{ g CO}_2\text{e}\cdot\text{MJ}^{-1}$ $[(-)45.7 \text{ to } (+)21.1]$ according to
409 energy-based allocation (**Table 2**). Thus, the CI of the median-CI locations is overestimated
410 approximately 27 (45Z), 69 (CFR), 80 (LCFS), and 99 (RFS) percent of the time by existing policy
411 frameworks (**Figure 1**).



412 **Figure 1.** Distribution of SAF CI based on field-specific LCA (using system expansion, hybrid allocation,
413 and energy allocation) compared to CIs of conventional jet fuel and calculated according to policy
414 specifications. The line and range for each policy-based CI are based on the policy-based CI sensitivity
415 analysis (**Section S1.1** and **S1.4** of the SI) with values for all ranges and lines reported in **Table S8** of the
416 SI.
417

418
419 The overall CI of SAF production is largely dependent on co-product electricity production
420 and the CI of miscanthus production (**Table 2**, **SI Table S2**). Thus, location-specific parameters,

421 namely the miscanthus CI and electricity CI, are important factors governing the overall SAF
 422 production CI (**Figure 2**). States that have both a low miscanthus CI and high electricity CI, such
 423 as South Dakota, are able to achieve the lowest total SAF CI (median of approximately -160 g
 424 CO₂e·MJ⁻¹) due to large amounts of soil carbon sequestration and displaced electricity from high-
 425 emissions sources. Conversely, higher total SAF CIs are likely to occur in states with higher
 426 (though still negative) miscanthus CIs and lower electricity CIs, such as New York. While varied
 427 electricity fuel sources are generally accounted for in low carbon fuel program fuel CI calculations,
 428 spatially varying feedstock CIs are generally ignored by the RFS and CFR. Given that spatially
 429 varying feedstock CIs can significantly affect the total fuel CI, they should be given appropriate
 430 consideration in low carbon fuel policies and tax credits.

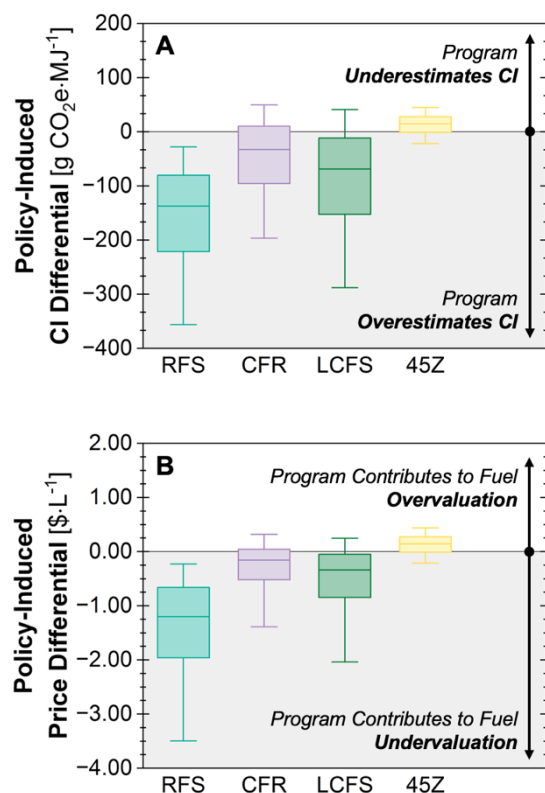


431
 432 **Figure 2.** SAF CI across miscanthus and electricity CIs. White diamond-shaped points indicate states for
 433 which miscanthus CIs were estimated by Fan *et al.* (2024). These points indicate the median CIs for each
 434 state based on simulations over 30 years at a 4 km resolution. Vertical clustering of states was the result
 435 of differences in electricity and natural gas emission factors which were based on balancing region and
 436 state, respectively.³⁵ Ranges of CIs for each state are detailed in **Table S8** of the SI.

437
 438 **Influence of Policy Specifications on SAF Characterization and Valuation**

439 Based on the CIs according to policies and spatially explicit LCA, the policy-induced CI differential
 440 for SAF produced from miscanthus is -137.4 g CO₂e·MJ⁻¹ [(-)356.3 to (-)27.8] for the RFS, -32.6

441 g CO₂e·MJ⁻¹ [(-)196.6 to (+)49.7] for the CFR, -68.9 g CO₂e·MJ⁻¹ [(-)287.7 to (+)40.8] for the LCFS,
 442 and +14.7 g CO₂e·MJ⁻¹ [(-)22.1 to (+)44.7] for the 45Z tax credit (**Figure 3A**). The overestimated
 443 CI on behalf of the RFS is explained by the non-producer-specific fuel pathway SAF CI used by
 444 the program (**Table 1**). The CFR overestimates and underestimates the CI similarly, but the use
 445 of a single program-based CI still cannot account for the potential variation in SAF CIs. The 45Z
 446 tax credit is the only policy that is likely to underestimate the SAF CI, because the CI is not
 447 producer-specific and estimates a higher amount of miscanthus-induced SOC storage (**Table 2**).



448
 449 **Figure 3.** Policy-induced (A) SAF CI differentials and (B) SAF price differentials. Negative differentials
 450 indicate that the SAF CI is overestimated and the fuel is undervalued, respectively, by the policy, and vice
 451 versa.

452
 453 Regardless of over- or underestimation, the wide range of policy-induced CI differentials
 454 highlights the limitations of using a single fuel CI for program participation (for all four policies, but
 455 particularly for the RFS and LCFS). The variation in the SAF CI estimated via spatially explicit
 456 LCA (which corresponds to the variation in CI differential in **Figure 3A**) can be attributed to
 457 multiple spatially varying and uncertain parameters along the biofuel supply chain (e.g.,
 458 miscanthus CI, biomass transport distance, biomass-to-biofuel conversion efficiency). Not

459 accounting for this variation among fuel producers and uncertainty inherent to fuel production
460 leads to falsely precise CI quantification and broader estimates of the benefits of biofuels. Properly
461 accounting for this variation and uncertainty (e.g., via field-specific LCA and explicit expression of
462 uncertainty) would lead to a more accurate quantification of carbon outcomes, properly incentivize
463 activities with the highest likelihood of reducing emissions, and help better achieve low carbon
464 fuel policy goals overall.

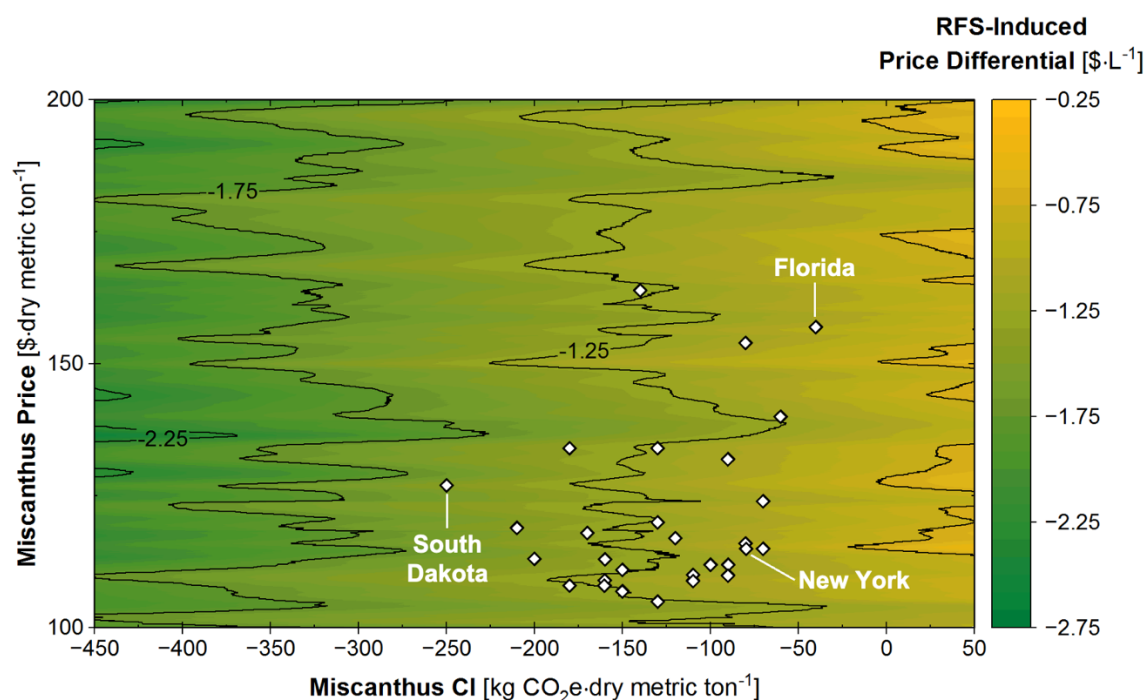
465 Based on policy-induced CI differentials, low carbon fuel program credit selling prices, and
466 the 45Z tax credit value, the policy-induced price differential for SAF produced from miscanthus
467 is $-1.20 \text{ \$}\cdot\text{L}^{-1}$ [(-)3.49 to (-)0.23] for the RFS, $-0.16 \text{ \$}\cdot\text{L}^{-1}$ [(-)1.39 to (+)0.32] for the CFR, $-0.34 \text{ \$}\cdot\text{L}^{-1}$
468 ¹ [(-)2.04 to (+)0.25] for the LCFS, and $+0.14 \text{ \$}\cdot\text{L}^{-1}$ [(-)0.22 to (+)0.44] for the 45Z tax credit (**Figure**
469 **3B**). Given that credit selling prices have been consistently greater than zero, the monetary
470 undervaluation of SAF is directly linked to overestimated CIs on behalf of low carbon fuel policies
471 (**Figure 3A**). These results suggest that the RFS will almost always undervalue SAF. While the
472 CFR-induced price differential was estimated based on LCFS credit prices, these results indicate
473 that the program is likely structured in a way that will undervalue SAF produced from miscanthus.
474 The 45Z tax credit is the only policy likely to overvalue SAF produced from miscanthus, due to
475 the underestimated CI on behalf of the policy and credit value.

476 The consistent undervaluation of a fuel produced from a cellulosic feedstock represents
477 one (out of many) challenges undercutting efforts to achieve cellulosic biofuel production targets
478 set by the RFS. If biofuel production credits are not more closely linked to actual CI reductions,
479 biofuel producers are less incentivized to undertake measures to reduce emissions. This is
480 particularly relevant for cellulosic biofuels, which are subject to high capital costs that cannot be
481 recovered with low fuel selling prices. Using field-specific LCA to more accurately quantify biofuel
482 CI reductions and allowing lower CI fuels to be sold at higher prices (because of higher priced
483 program credits) would make more money available to low-GHG technologies (including
484 herbaceous feedstocks) and measurement tools to facilitate field-specific LCA.

485 The overall minimum selling price of SAF was estimated as $2.77 \text{ \$}\cdot\text{L}^{-1}$ [2.23 to 3.62], and
486 is largely dependent on capital costs and the miscanthus price (SI **Table S2**). Thus, location-
487 specific parameters such as the miscanthus price and location capital cost factor (LCCF, an
488 estimate of how capital costs vary across the U.S.) are important predictors of the SAF selling
489 price across locations (SI **Figure S2**). SAF produced from miscanthus with the lowest CI (e.g.,
490 miscanthus produced in South Dakota) will be the most undervalued by the RFS (an

491 undervaluation of 1.35 $\text{\$}\cdot\text{L}^{-1}$ [0.31 to 3.67]; **Figure 4**, SI **Table S4**). SAF produced from the most
492 expensive miscanthus (e.g., miscanthus produced in Florida) is less undervalued (i.e., more
493 accurately valued, an undervaluation of 0.80 $\text{\$}\cdot\text{L}^{-1}$ [0.23 to 2.66]) because the higher production
494 cost offsets the effect of the CI overestimation and corresponding fuel undervaluation.

495 An ATJ SAF biorefinery operating in South Dakota could achieve a minimum SAF selling
496 price of 2.57 $\text{\$}\cdot\text{L}^{-1}$ [2.18 to 3.14] (SI **Figure S2**, SI **Table S4**). If the RFS were to more accurately
497 quantify the CI of this SAF (i.e., via field-specific LCA) and the policy induced price differential of
498 1.35 $\text{\$}\cdot\text{L}^{-1}$ (**Figure 4**) were eliminated, the biorefinery's required SAF selling price would instead
499 be 1.22 $\text{\$}\cdot\text{L}^{-1}$. Factors that affect farm-specific biofuel CIs should be considered in low carbon fuel
500 policies to better promote favorable farming locations (i.e., locations that can achieve low fuel CIs)
501 that might be subject to higher production costs, such as South Dakota.



502
503 **Figure 4.** RFS-induced SAF price differential across miscanthus CI and miscanthus price. White diamond-
504 shaped points indicate states for which miscanthus CIs and prices were estimated by Fan et al. (2024) and
505 Lee et al. (2023), respectively. These points indicate the median CI and price for each state. Ranges of CIs
506 and prices for each state are detailed in SI **Table S8**.
507

508 Feasibility and Need for Innovation

509 *Technological Capabilities*

510 While miscanthus has the potential to contribute to the U.S. bioeconomy,⁵⁷ this crop has not yet
511 been widely cultivated in the U.S. (e.g., in 2022, only approximately 68,000 metric tons were
512 produced across 9 states,⁵⁸ whereas approximately 343 million metric tons of corn grain were
513 produced across 49 states⁵⁹). The factors influencing the decision by individual farmers to
514 cultivate novel energy crops such as miscanthus have been studied extensively (e.g.,^{60–63}). While
515 accurate quantification of biomass feedstock and biofuel CIs is important, achieving the benefits
516 of SOC sequestration and low carbon biofuels will also depend on farmers' transition to growing
517 herbaceous energy crops. Cost subsidies for crop establishment⁶¹ and payments for carbon
518 mitigation⁶³ could make these novel crops more appealing to farmers and facilitate this transition.

519 Much of the technology needed to collect measurements used to quantify field-specific
520 feedstock CIs is well developed. The cost of deploying the technology could be an impediment,
521 however. For example, an eddy-covariance flux tower used to collect carbon flux data associated
522 with biomass growth costs approximately \$100,000 to install and operate.¹⁴ If technology used to
523 perform field-specific LCA technology is implemented, however, it seems likely that deployment
524 costs could be recovered relatively quickly (i.e., assuming that current fuel price differentials
525 highlighted in **Figure 4B** are negated by improved CI quantification and fuels are sold for higher
526 prices). Further, rather than intensive monitoring on every field, a system-of-systems approach
527 that combines in-field measurements, biogeochemical modeling, and big data analytics could
528 enable better and more cost effective quantification of field-level carbon outcomes;¹⁴ however,
529 further financial investment is needed to continue to develop and prototype MMRV technologies.

530 The software generally used for biofuel CI calculations (e.g., GREET, openLCA) are
531 robust LCA tools that are capable of incorporating field-specific factors and accounting for some
532 sources of uncertainty. In general, policy provisions and LCA tools already allow for producer-
533 specific inputs when modeling biorefinery operations (e.g., amounts of ancillary chemicals
534 including enzymes, acids, caustic; region-specific electricity fuel sources). Given the capabilities
535 of these LCA tools, this scope of consideration could easily be expanded to field-specific inputs
536 (e.g., fertilizers, fuel) and biogeochemistry (e.g., SOC dynamics, given appropriate
537 measurements and monitoring take place). The models themselves are also able to account for
538 uncertainty (in the amounts and CIs of chemical inputs, etc.), but none of the low carbon fuel
539 policies seem immediately capable to do so. Program administrators should consider the benefits

540 of accounting for uncertainty in CI calculations compared to assigning false precision to carbon
541 outcomes via single-value CI estimates.

542 Regulatory and Legal Precedent

543 The results of this study add to a body of literature (e.g., ^{7,42,64}) suggesting that the RFS technology
544 mandate does not properly incentivize the production of low carbon fuels, while expanding the
545 analysis to additional policies and a proposed pathway from a perennial grass to SAF. Field-
546 specific LCA could better incentivize actions to meaningfully lower crop and biofuel carbon
547 intensity and, in doing so, provide the capital necessary to implement in-field monitoring. The CFR
548 is a more recently enacted performance-based program that appears better structured to
549 characterize and value cellulosic biofuels (**Figure 4**). In anticipation of the continued
550 establishment of RFS annual volume requirements, program administrators could consider
551 restructuring the program to a performance-based standard to promote climate and financial
552 benefits.

553 The Carbon Sequestration and Collaboration Act (CSCA),⁶⁵ while not yet enacted into law,
554 indicates a desire on behalf of the U.S. Congress to better understand carbon sequestration
555 methods. The CSCA specifically acknowledges the need to understand the variation in carbon
556 storage potential in agricultural soil (as well as forests and geologic formations) across
557 geographies. If enacted, the CSCA will direct the Department of Energy, the Department of the
558 Interior, and the Department of Agriculture to establish carbon sequestration research initiatives,
559 which could aid the development of technologies necessary to perform field-specific LCA.

560 Path Forward

561 While cellulosic biofuels present a viable option to lower GHG emissions associated with aviation,
562 current low carbon fuel program provisions meant to incentivize their development are ineffective.
563 Structural undervaluation of fuels that achieve lower carbon intensities in turn reduces the amount
564 of money available to further develop these technologies. While the LCFS and CFR structurally
565 promote the production of fuels that achieve the lowest carbon intensities (i.e., they generate
566 program credits based on CI reductions), program methodological requirements may still lead to
567 the mischaracterization and undervaluation of these fuels. Due to its CI threshold-based structure,
568 the RFS does not properly incentivize the production of fuels that achieve lower CIs. Advanced
569 energy feedstocks such as miscanthus have the potential to significantly lower the CI of SAF
570 production, but this potential is not properly accounted for by existing program provisions. For all
571 three low carbon fuel policies, issues with fuel characterization could be resolved by allowing fuel

572 producers to calculate total fuel CIs that account for field-specific biomass production. Field-
573 specific LCA presents a practicable means to more accurately quantify the outcomes of biofuel
574 production and help society realize the benefits of cellulosic biofuels.

575

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581 reflect the views of the U.S. Department of Energy.

582

583 **Supporting Information Available**

584 The Supporting Information includes:

585 Detailed policy-specific CI calculations, description of input parameter distributions and
586 sensitivity analysis results, description of low carbon fuel program credit prices, expanded
587 TEA results.

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