

1 THE VARIABILITY OF VOLATILE ORGANIC COMPOUNDS IN CLINICAL ENVIRONMENTS.

2 Dahlia Salman^a, Wadah Ibrahim^{b,c}, Amisha Kanabar^a, Bo Zhao^{b,c}, Amisha Singapuri^{b,c}, Michael
3 Wilde^d, Rebecca L Cordell^d, Teresa McNally^b, Dorota Ruszkiewicz^a, Andria Hadjithekli^a, Robert Free^{b,c},
4 Neil Greening^{b,c}, Erol A Gaillard^b, Caroline Beardsmore^b, Paul Monks^d, Chris Brightling^{b,c}, Salman
5 Siddiqui^{b,c}, C L Paul Thomas^{a*} On behalf of the EMBER consortium

6 ***Corresponding author:** C.L.P.Thomas@lboro.ac.uk

7 a Department of Chemistry, Loughborough University, Loughborough, LE11 3TU, UK.

8 b College of Life Sciences, Department of Respiratory Sciences, University of Leicester,
9 University Road, Leicester LE1 7RH

10 c Leicester NIHR Biomedical Research Centre (Respiratory theme), Glenfield Hospital, Groby
11 Road, Leicester LE3 9QP

12 d Department of Chemistry, University of Leicester, University Road, Leicester, LE1 7RH

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14 **KEYWORDS:**

15 Volatile organic compounds, environmental VOC, indoor air pollution, breath analysis, TD-

16 GC-MS; thermal desorption-gas chromatography-mass spectrometry.

17

18 **ABSTRACT**

19 The development of clinical breath-analysis is confounded by the variability of background
20 volatile organic compounds (VOC). Reliable interpretation of clinical breath-analysis at
21 individual, and cohort levels requires characterisation of clinical-VOC levels and exposures.
22 Active-sampling with thermal-desorption/gas chromatography-mass spectrometry recorded
23 and evaluated VOC concentrations in 245 samples of indoor air from three sites in a large
24 NHS provider trust in the UK over 27 months.

25 Data deconvolution, alignment and clustering isolated 7344 features attributable to VOC
26 and mapped the variability (composition and concentration) of respirable clinical VOC. 328
27 VOC were observed in more than 5% of the samples and 68 VOC appeared in more than
28 30% of samples. Common VOC were associated with exogenous and endogenous sources
29 and 17 VOC were identified as seasonal differentiators. The presence of metabolites from
30 the anaesthetic sevoflurane, and putative-disease biomarkers in room air, indicated that
31 exhaled VOC were a source of background-pollution in clinical breath-testing activity.

32 With the exception of solvents, and PPE waxes, exhaled VOC concentrations above $3 \mu\text{g m}^{-3}$
33 are unlikely to arise from room air contamination, and in the absence of extensive survey-
34 data, this level could be applied as a threshold for inclusion in studies, removing a potential
35 environmental confounding-factor in developing breath-based diagnostics.

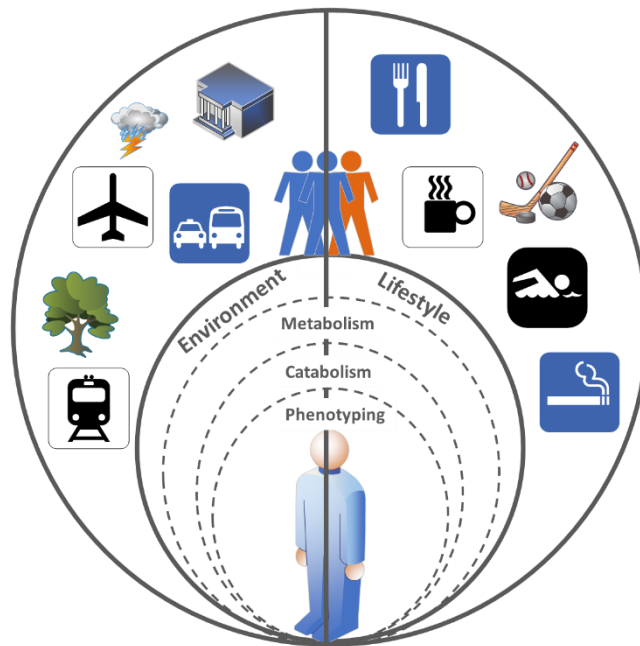
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37 INTRODUCTION

38 Breathomics, is being developed to stratify patient phenotypes, and monitor metabolic and
39 disease mechanisms¹⁻³, and biomarker discovery with breathomics is being applied to
40 conditions that include respiratory disease, cancer, infections and pulmonary illnesses⁴⁻⁶.
41 Further, breath analysis for assessing occupational exposure to VOC is well-established⁷⁻⁹.
42 A challenge in the development of clinical breath-testing is accounting for the heterogeneity
43 of patient responses to variable backgrounds of environmental VOC. Failure to adequately
44 address this factor may confound breathomic biomarker discovery and breath-testing
45 activity ^{10,11}.

46 Variability in breath biochemistry derives from: environmental contaminants; genetics; diet
47 and lifestyle; diurnal changes in metabolism; endocrine cycles; medication; emotional/
48 psychological states; and, disease progression, see Figure 1. Environmental VOC exposure
49 from inhalation, trans-dermal absorption or ingestion may result in elevated exhaled
50 concentrations of VOC, and/or metabolic/catabolic products not originally in the
51 environment¹². Further, endogenous VOC, and disease markers may also be obscured,
52 consequently, the VOC exposome generates a risk of false-attribution leading to breath-
53 testing outcomes that are difficult to reproduce or translate into clinical practice.

54 Studies of VOC in hospitals, homes, and workplace settings over a prolonged period have
55 shown exhaled breath VOC and environmental VOC often contain the same VOC. Common
56 VOC have been observed in matched samples from ventilators, blood, breath and the room-
57 air of clinics^{13,14}. Widely reported VOC biomarkers for respiratory diseases, such as
58 cyclohexanone for COPD¹⁵ have also been detected in indoor air at concentrations ($1.13 \mu\text{g m}^{-3}$
59 m^{-3})¹⁶, close to those in exhaled breath (0.4 to $10 \mu\text{g m}^{-3}$)¹⁷.



60

61 Figure 1 The observed exposome for each individual includes external factors such as
 62 environmental contamination and lifestyle, and also internal factors such as
 63 metabolism, catabolism and differences in phenotype. These are all factors in
 64 exhaled breath VOC.

65 Spatial and temporal variability of VOC in clinical settings has been observed with acetone,
 66 ethanol and propanol concentrations found to vary significantly while other VOC did not.
 67 Within the same study exhaled concentrations of acetone, ethanol, acetic acid, ammonia,
 68 isoprene and hydrogen cyanide were found to be higher in the breath of 10 clinical staff
 69 than in their surrounding environment, and propanol (a disinfectant) was at higher
 70 environmental concentrations¹⁸.

71 The sources and patterns of VOC in the indoor air of dwellings have been studied extensively
 72 with 2246 samples monitoring a panel of 61 VOC, enabling seasonal effects to be attributed
 73 to 8 proposed VOC patterns and sources ¹⁶.

74 It is also helpful to note that the exogenous VOC toluene and benzene may also arise from
 75 degradation of the adsorbent Tenax TA (used in thermal desorption tubes), highlighting the
 76 point that analytical systems also create a trace VOC profile that may vary in response to
 77 changes in the environmental background (acidity, basicity, ozone and humidity), and that
 78 Tenax and multibed thermal desorption sampling tubes should not be assumed to be inert¹⁹.

79 Exogenous VOC may affect breath marker discovery by: raising exhaled breath
80 concentrations (false-positive); raising the concentration threshold for inclusion of an
81 exhaled VOC as coming from endogenous origin causing endogenous compounds to be
82 excluded from a study (false-negative); give rise to catabolite signals with endogenous
83 features (false-positive); and/or act as a contrast agent that leads to correct identification
84 of a process, mechanism, or breath-biochemical derangement, but in a non-reproducible
85 manner.

86 This study addressed the environmental VOC breathomics issue, and collected room-air
87 samples from three sites in a large NHS provider trust in the UK – University Hospitals of
88 Leicester. The aim was to characterise the variability of composition and concentration of
89 clinical-VOC, and establish , thresholds, or reference-levels, to inform occupational-exposure
90 and breathomics workflows²⁰. The resultant sample set captures real-world operational
91 clinical-VOC exposure across 27 months and reveals the variability and extent of exogenous
92 and endogenous VOC present in clinical room air at trace levels .

93 **MATERIALS AND METHODS**

94 This study and sampling campaign was part of a larger prospective, real-world, observational
95 study, carried out in a tertiary cardio-respiratory centre in Leicester, United Kingdom, and .
96 the study design, and methodology have been described in detail previously²⁰.

97 Two hundred and forty five room-air samples were collected, of which 225 samples were
98 collected from Respiratory Medicine and Clinical Decision Units at Glenfield Hospital,
99 Leicester, UK and 20 samples were collected from Paediatric Respiratory Medicine,
100 Leicester Royal Infirmary Hospital, Leicester, UK, from November 2016 to February 2019.
101 These samples were taken during normal clinical operations in the presence of patients
102 participating in respiratory research, and research clinical staff. On occasions, other patients
103 and staff were present, particularly if the samples were collected from a ward-bay.

104 VOC artefacts and contamination were reduced/eliminated from the materials and
105 components used within this study with appropriate combinations of solvent cleaning,
106 vacuum polishing, and temperature conditioning. All equipment was sealed in aluminium
107 packaging ready for use before shipping to the clinic. Research and clinical staff involved in
108 the study were trained to follow specifically designed sampling standard operating protocols,
109 and were proficiency tested in their use²⁴. Sampling protocol checklists were used to verify
110 compliance with the standard operating protocols.

111 Indoor air was sampled (1000 cm³) using an Escort ELF pump (Part No. 497702, MSA), at
112 a flow rate of 500 cm³ min⁻¹ for 120 s onto a Tenax®/Carbotrap 1TD hydrophobic adsorbent
113 tube (Part No. C2-AXXX-5032, Markes International Ltd, Llantrisant, UK, see Figure S1).
114 Samples were sealed and immediately stored at ca. 4 °C, before shipping to Loughborough
115 Centre for Analytical Science within approximately 3 days. Environmental samples were dry-
116 purged as soon as possible upon receipt with a 120 cm³ of purified nitrogen at a flow rate
117 of 60 cm³ min⁻¹. Toluene-D8 (69 pg) and trichloromethane-d (280 pg) internal standards
118 were loaded during the dry purge onto the sample tube with a six-port valve attached to a
119 permeation tube-based test atmosphere generator (constructed in -house). Dry-purged and
120 internal-standard spiked samples were then sealed and stored at 4 °C prior to analysis.

121 TD-GC-MS operating conditions

122 Samples were analysed by thermal-desorption/gas-chromatography/mass-spectrometry
123 (TD/GC/qMS). A Unity-2 thermal desorption unit (Markes International, Cardiff, UK) was
124 interfaced to a GC (Agilent, 7890A) coupled to a quadrupole mass spectrometer (Agilent MS
125 5977A). The VOC collected during sampling were recovered and concentrated into a
126 hydrophobic cold trap yielding a 10,000-fold enrichment, the instrumentation parameters
127 are summarised in Table S1.

128 Statistical Process Control

129 Instrumentation performance was continuously monitored by analysing 0.2 µl of a reference
130 mixture containing 20 standards (Table S2) daily before analysis. Instrument performance
131 was evaluated by monitoring the *Z*-scores of: retention time, peak area, height, width, and
132 symmetry for the 20 standards in the reference mixture. Analysis was undertaken when
133 instruments were within $Z = \pm 3$ for more than 80% of the 100 quality control parameters²².
134 The two internal standards were also monitored to track the combined stability of the TD-
135 GC-MS analysis and dry purging system.

136 Data processing

137 The TD-GC-MS data were deconvolved and an average of 120 VOC features per sample
138 (AnalyzerPro Spectral Works, UK) were extracted. The deconvolution method was optimised
139 to minimise over-deconvolution, (AnalyzerPro software method parameters were: minimum
140 peak area value, = 200, S/N= 3, width of peak = 0.01 and smoothing factor=3). The
141 extracted features were aligned using Kovats retention indexing (AnalyzerPro Spectral
142 Works, UK) ²³, and clustered using the VOCcluster algorithm²⁴ that assigned a unique
143 identifier in the form of (ERI- m/z_1 - m/z_2 - m/z_3 - m/z_4 - m/z_5) to each VOC isolated and
144 grouped from the 245 samples; ERI indicated the retention index for the VOC environmental-
145 feature and m/z_1 ... m/z_n were the nominal masses of the compound's ion fragments in
146 decreasing order of abundance needed to uniquely define the deconvolved VOC features
147 within the data-cube. The resultant sorted, and grouped features were consolidated into an
148 environmental-VOC-data matrix that contained the extracted peak areas for each of the
149 features isolated from each room-air sample.

150 Ubiquitous siloxanes arising from analytical artefacts were also removed from the
151 Environmental-VOC-Matrix, and 7 samples were discarded as they did not meet the quality
152 control measures.

153 **Multivariate analysis for seasonality effects**

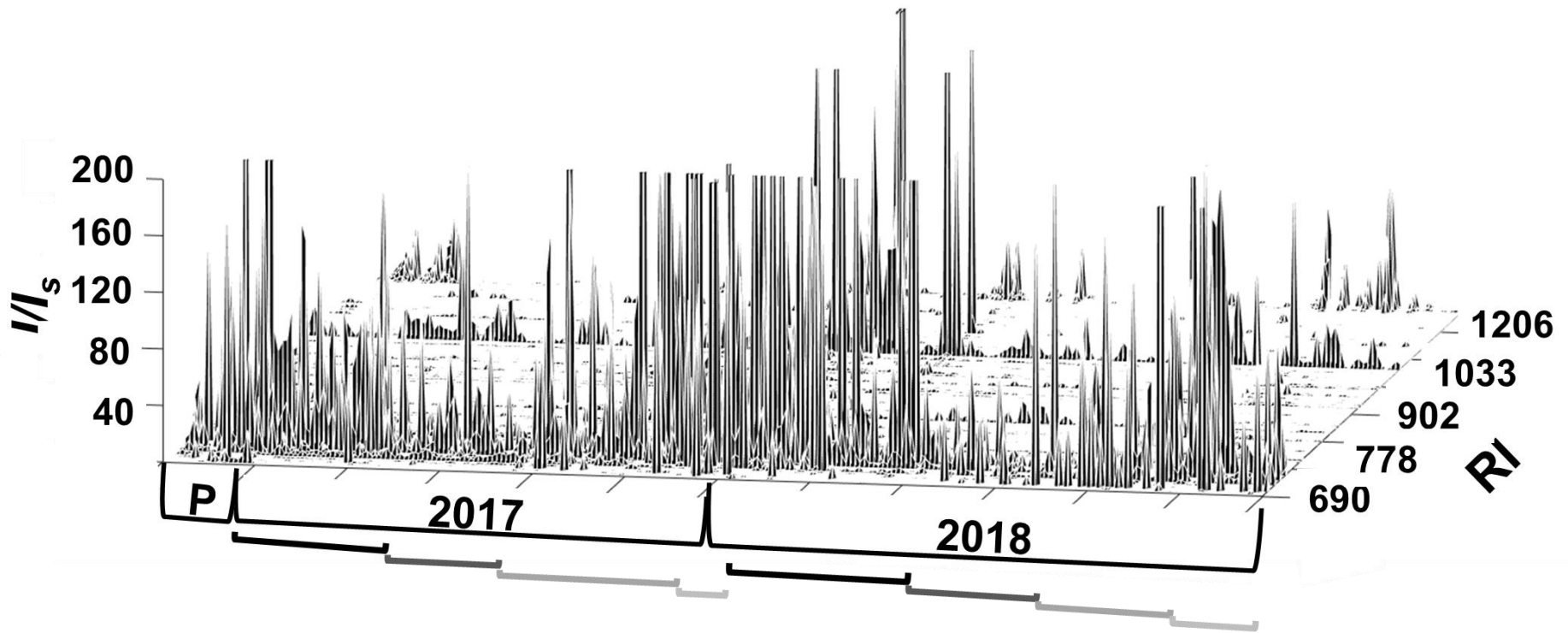
154 Once compounds that occurred in less than 30 % of samples had been excluded, the matrix
155 features were \log_{10} transformed and Pareto scaled before multivariate analysis was used to
156 determine if there were any seasonal effects in the study ^{25,26}. The data was classified into
157 two groups; September-February, n=99 samples; and, March-August, n=139 samples.
158 Orthogonal partial least squares-discriminant analysis (OPLS-DA) was initially used to
159 identify seasonally invariant VOC, and these were removed from the analysis leaving 44
160 compounds for unsupervised principle components analysis (PCA). This approach followed
161 a multivariate statistical processing workflow using SIMCA-P+ software with integrated 7-
162 fold cross-validation to protect against overfitting (Version 16.1, Umetrics, UK)²⁵.
163 Discriminating compounds were putatively assigned a Level-2 identification in accordance
164 with the Metabolomics Standards Initiative²⁷, based on retention-index and NIST mass-
165 spectral library matches.

166 **RESULTS**

167 Two hundred and forty five clinical room air samples yielded 7344 VOC features, with 328
168 compounds that occurred more than 12 times (5%) that were selected for further
169 assessment, see Figure 2. Sixty eight compounds appeared in more than 30% of the total
170 samples. Thirty nine of these have been assigned a putative identity (Class 2 identification
171 level²⁷) and a further 11 compounds were classified as hydrocarbons (Class 3 identification
172 level²⁷), see Table 1. These data were used for multivariate analysis. (The identities of
173 eighteen of these compounds have yet to be elucidated.)

174 **Seasonality**

175 Seasonal variations in concentration, were identified from multi-variate analysis with 17
176 compounds found to be present at higher concentrations in samples obtained between
177 September and February, Figure S2 and Table S2. In 12 of the 17 compounds the
178 differences in concentration were statistically significantly different (normalised to the
179 internal standard).



180

181 Figure 2: Common indoor VOC (> 5% of samples) found in air samples from adult and the paediatrics clinics taken during November 2016
 182 to February 2019. The VOC are plotted along the sampling timeline against their retention index (RI) values with their intensity
 183 normalised to 280 pg of C(2H)Cl₃. P represents the data from the children's clinic. The rest of the data were obtained from the
 184 adults' wards with seasons of each year (spring, summer, autumn and winter) indicated by the grey shaded brackets found
 185 underneath each year. Note that 7016 VOC isolated from less than 5% of samples are not included.

186

187 Table 1 50 VOC that have been putatively identified that appeared in more than 30% of the samples. ERI code (showing the retention
 188 index and quantifier m/z 1 and the qualifier ions m/z 2-5), NIST-library matched identification, CAS number, mean intensity ratio
 189 to C(2H)Cl³ internal standard (\bar{I}/I_{IS}), median intensity ratio to C(2H)Cl₃ internal standard (\tilde{I}/I_{IS}), standard deviation (*S*) and
 190 frequency of observation (*F*)

VOC no.	ERI Code generated by VOCcluster ²⁴	Putative compound identification	CAS	\bar{I}/I_{IS}	\tilde{I}/I_{IS}	<i>S</i>	<i>F</i>
1	ERI-670-101-151-103-85-153	•1,1,2-trichloro-1,2,2-trifluoroethane-	76-13-1	0.64	0.72	0.32	91
2	ERI-683-45-46-43-42-44	^b ethanol	64-17-5	21.9	11.37	39.72	234
3	ERI-685-45-43-58-42-0	2-propanol	67-63-0	104.1	38.95	218.3	169
4	ERI-688-67-53-68-51-50	•2-methyl-1,3-butadiene	78-79-5	0.64	0.44	0.84	144
5	ERI-700-43-45-60-42-41	^b ethanoic acid	64-19-7	2.77	1.72	3.41	184
6	ERI-712-78-77-51-50-52	^b benzene	71-43-2	1.76	1.44	2.68	220
7	ERI-778-91-92-65-63-51	^b toluene	108-88-3	1.71	0.97	2.71	204
8	ERI-787-41-86-45-68-68	2-methyl-2-propenoic acid,	79-41-4	0.42	0.33	0.26	80
9	ERI-791-50-46-66-66-82	N-containing compound		0.46	0.52	0.19	83
10	ERI-801-43-41-56-43-57	^b hexanal	66-25-1	0.40	0.29	0.30	157
11	ERI-814-43-56-41-73-61	butyl ethanoate	123-86-4	0.50	0.38	0.48	113
12	ERI-854-43-72-57-42-58	^b 2-butanone	78-93-3	0.39	0.28	0.27	93
13	ERI-861-91-106-51-65-65	^b ethyl benzene	100-41-4	0.50	0.30	0.51	161
14	ERI-869-91-106-105-77-51	^b 1,4-dimethylbenzene	106-42-3	1.62	0.52	5.27	134
15	ERI-893-55-42-98-98-69	^b cyclohexanone	108-94-1	0.61	0.49	0.71	94
16	ERI-895-91-106-105-77-51	^b 1,2-dimethylbenzene,	95-47-6	0.64	0.40	1.50	170
17	ERI-907-57-45-41-87-56	^b 2-butoxy-ethanol,	111-76-2	0.55	0.37	0.71	80
18	ERI-935-93-91-92-77-79	(1R)-2,6,6-trimethylbicyclo[3.1.1]hept-2-ene	7785-70-8	0.57	0.47	0.51	157
19	ERI-960-77-105-106-51-50	^b benzaldehyde	100-52-7	1.05	0.63	2.44	231
20	ERI-983-94-66-65-63-55	^b phenol	13127-88-3	1.09	0.64	2.86	228
21	ERI-985-103-76-50-75-75	•trans-cinnamic acid ((2E)-3-Phenyl-2-propenoate)	140-10-3	0.37	0.25	0.31	144
22	ERI-987-43-41-55-69-108	^b 6-methyl-5-hepten-2-one,	110-93-0	0.48	0.39	0.37	195
23	ERI-995-105-120-91-103-51	^b 1-methylethylbenzene	98-82-8	0.50	0.34	0.97	80
24	ERI-997-105-120-77-91-119	^b 1,2,3-trimethylbenzene,	526-73-8	0.54	0.49	0.42	77
25	ERI-1003-41-43-57-56-0	^b octanal	124-13-0	0.39	0.29	0.27	110
26	ERI-1026-119-134-134-117-77	1-methyl-4-(1-methylethyl) benzene	99-87-6	0.52	0.36	0.60	101
27	ERI-1029-57-41-43-55-56	^b 2-ethyl 1-hexanol-	104-76-7	1.21	0.68	1.86	222
28	ERI-1030-68-67-93-79-53	^b 1-Methyl-4-methylethenylcyclohexene	138-86-3	2.15	0.57	13.46	212
29	ERI-1035-79-108-77-107-51	^b phenylmethanol	100-51-6	0.46	0.38	0.31	100
30	ERI-1068-105-77-51-120-50	^b 1-phenylethanone	98-86-2	0.50	0.34	0.46	172
31	ERI-1073-59-43-55-41-67	2,6-dimethyl -7-octen-2-ol	18479-58-8	0.76	0.38	1.80	111

32	ERI-1104-57-41-43-55-56	^b nonanal	124-19-6	0.79	0.64	0.69	215
33	ERI-1145-55-41-97-70-57	[•] 6-methyl-1-octanol	110453-78-6	0.46	0.38	0.43	84
34	ERI-1169-105-77-122-51-50	N-containing compound	N/A	0.39	0.31	0.27	82
35	ERI-1176-71-81-95-41-55	(1R,2R,5R)-2-Isopropyl-5-methylcyclohexanol	491-02-1	0.95	0.45	3.32	89
36	ERI-1206-41-43-57-55-44	^b decanal	112-31-2	0.43	0.32	0.32	155
37	ERI-1377-71-43-56-89-41	3-hydroxy-2,4,4-trimethylpentyl 2-methylpropanoate	74367-34-3	0.59	0.33	0.93	92
38	ERI-1606-71-43-41-56-55	2,4,4-trimethyl-1,3-pentanediy l bis (2-methyl propanoate)	74381-40-1	2.02	0.53	3.25	112
39	ERI-2333-102-260-232-231-76	2,5-diphenyl-2,5-cyclohexadiene-1,4-dione,-	844-51-9	3.10	0.58	8.77	74
40	ERI-679-44-43-41-42-56			6.75	2.55	21.55	219
41	ERI-690-43-57-41-56-42			0.76	0.41	1.41	119
42	ERI-722-43-57-56-58-86			0.42	0.32	0.26	98
43	ERI-991-57-56-41-43-55			0.72	0.39	1.39	161
44	ERI-1000-43-57-41-71-85			0.61	0.34	2.21	173
45	ERI-1100-57-43-41-71-55	hydrocarbon		0.65	0.40	1.67	86
46	ERI-1200-57-43-41-71-85			0.42	0.30	0.37	192
47	ERI-1303-57-43-41-67-85			0.46	0.38	0.35	86
48	ERI-1400-57-43-71-41-85			0.45	0.34	0.38	159
49	ERI-1605-57-43-71-41-85			0.62	0.40	0.85	96
50	ERI-2342-57-43-71-85-41			14.58	0.94	64.72	74

191 [•]: qualified retention index match,

192 ^b: previously observed in breath

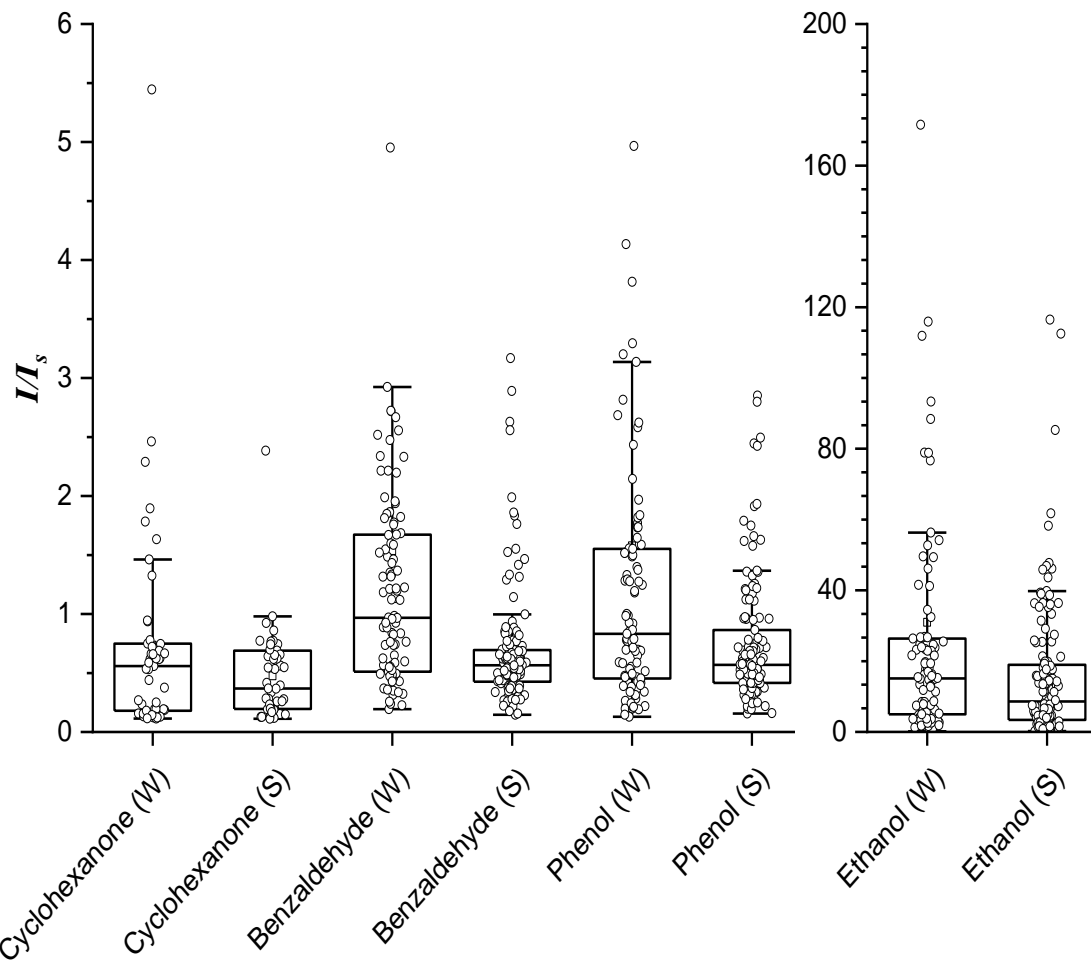
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194 Cyclohexanone (1.5-fold increase, t- statistic = 1.79, one-tailed critical t-value = 1.66, p =
195 0.038 with 62 degrees of freedom) has been previously reported as a marker for COPD
196 (sensitivity = 60% and specificity = 91%)²⁸. Benzaldehyde (2.2-fold increase t-statistic =
197 2.27, one-tailed critical t-value = 1.66, p=0.012 with 100 degrees of freedom) is involved in
198 fatty-acid and tryptophan metabolism as well as glycolysis/gluconeogenesis²⁹. Exhaled
199 benzaldehyde has also been reported: in breath samples from participants with severe
200 pulmonary arterial hypertension, as well as healthy individuals³⁰; possibly generated from
201 bacterial degradation of common amino acids such as phenylalanine, tryptophan or
202 tyrosine³⁰; lung cancer cell lines (n=6) discriminating from healthy control cell lines (n=1) ³¹;
203 and, has also been proposed to originate from exposure to tobacco smoke, radiation or air
204 pollution with peroxidative properties capable of damaging enzymes and DNA ³². Phenol
205 (2.1-fold increase, t-value statistic = 1.84, one-tailed critical t-value = 1.66, p=0.034 with
206 95 degrees of freedom.) is associated with petroleum products as well as tobacco smoke³³.
207 Phenol is also associated with oesophageal or gastric adenocarcinoma, and has been
208 observed to be significantly higher in cancer patients (n=81) compared to healthy individuals
209 (n=129) (P<0.05)³⁴, Figure 3.

210 Ethanol (2-fold increase, t- statistic = 2.63, one-tailed critical t-value = 1.66, p = 0.0048 with
211 113 degrees of freedom) and ethanal (2-fold increase but not statistically significant) may
212 be attributed to ethanol consumption and metabolism with the observed increases in
213 abundances of due to higher seasonal alcohol consumption. (Note that the ethanal mass
214 spectrum did not fall completely in the mass spectrometric scan range and verification with
215 derivatisation was not undertaken.)

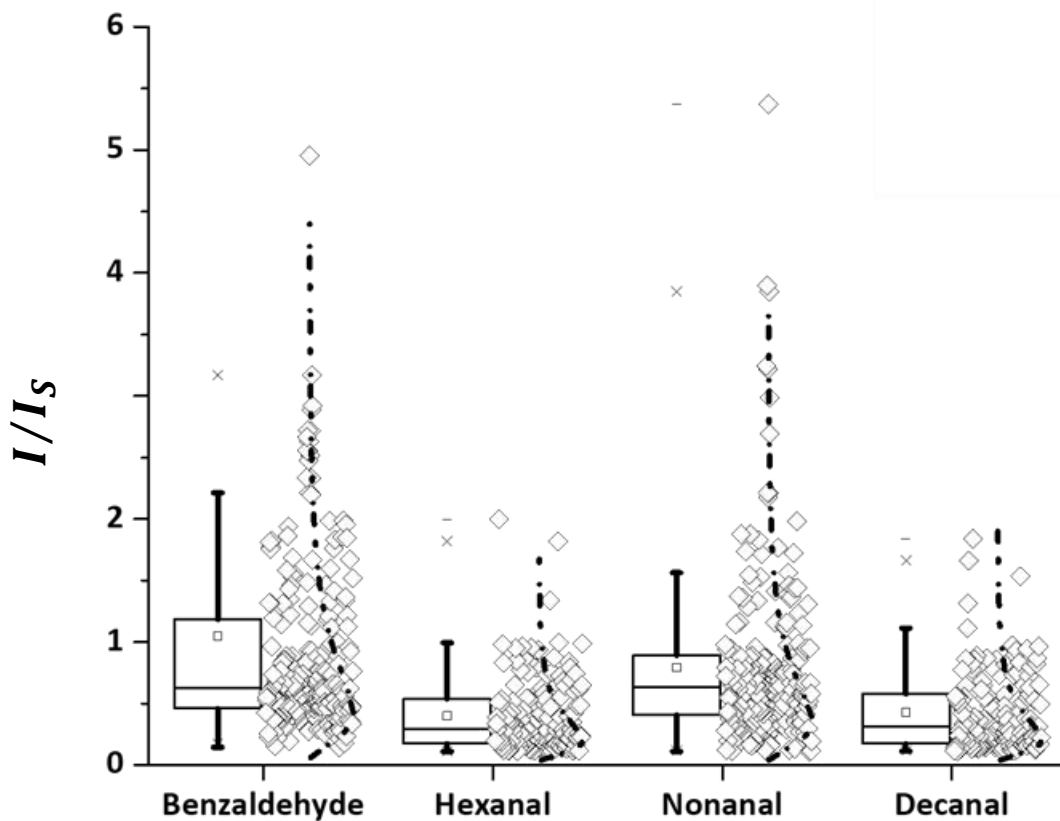
216 Cleaning product/disinfectants such as 2-propanol (2-fold increase, t- statistic = 2.18, one-
217 tailed critical t-value = 1.66, p = 0.016 with 85 degrees of freedom) were also higher in
218 winter

219 Variation between different clinical settings was discernible with 3 compounds unique to the
 220 paediatric setting that were related to the paediatric anaesthetic sevoflurane ^{35,36}.
 221 Sevoflurane and its two metabolites 1,1,1,3,3,3-hexafluoro-2-propanone, and 1,1,1,3,3,3-
 222 hexafluoro-2-propanol were putatively identified based on NIST-mass-spectral library
 223 matches, see Figure S3. Sevoflurane and its metabolite 2-propanol, 1,1,1,3,3,3-hexafluoro
 224 have been measured previously in human breath (n=6), and used to build a three-
 225 compartment pharmacokinetics model to study environmental contaminants and breath
 226 data³⁷. The presence of these VOC may well have arisen from a ventilation circuit shared
 227 with a surgical theatre on a different floor of the building.



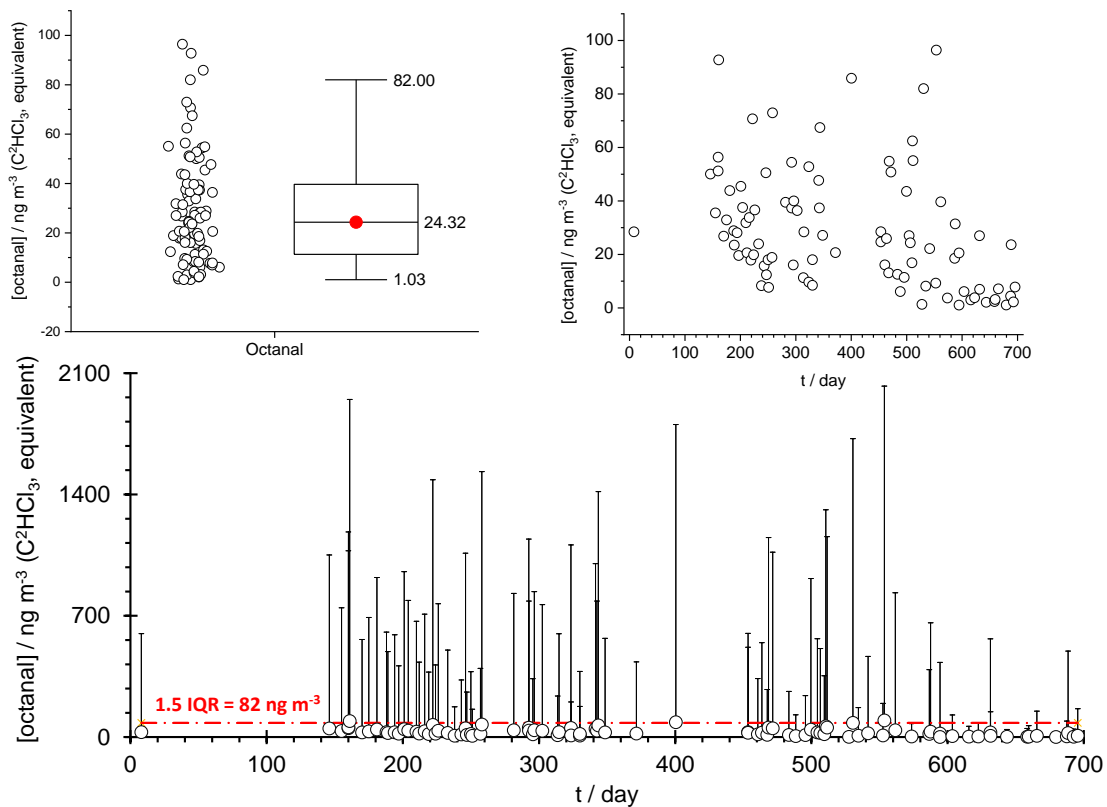
228
 229 Figure 3 Box whisker plots for cyclohexanone, benzaldehyde, phenol, and ethanol
 230 contrasting the differences in profiles between Winter (W), September-February,
 231 n=99 samples and Summer (s), March-August, n=139 samples. Note that winter
 232 outliers for benzaldehyde and phenol with I/I_s values of 36.6 and 42.6

233 respectively are not shown. I/I_s denotes signal intensity relative to the internal
 234 standard $C(^2H)Cl_3$ signal.
 235 As well as cyclohexanone and benzaldehyde other disease markers were found in the room
 236 air of clinics and Figure 4 also includes the distributions of the inflammatory biomarkers
 237 nonanal, hexanal and decanal across the sampling campaign.



238
 239 Figure 4 Box plots and lognormal distribution curves for benzaldehyde, hexanal, nonanal
 240 and 2-propanol, decanal, showing the prevalence of these often-reported breath
 241 biomarkers in clinical room air samples.
 242 Another commonly reported breath biomarker, octanal, was also present, and provides a
 243 useful case-study into the variability of the concentrations encountered and the frequency
 244 of occurrence. Octanal was isolated from 101 environmental samples, a frequency of
 245 occurrence of 42%, the highest concentration observed was estimated to be 96 ng m^{-3}
 246 (expressed as a $C(^2H)Cl_3$ equivalent) with a minimum observed concentration of 1.04 ng m^{-3}
 247 ³, and the median observed octanal concentration was 24.32 ng m^{-3} . Tests for normality and
 248 log-normality (Shapiro-Wilk) indicated a non-normal distribution and assessment of the time

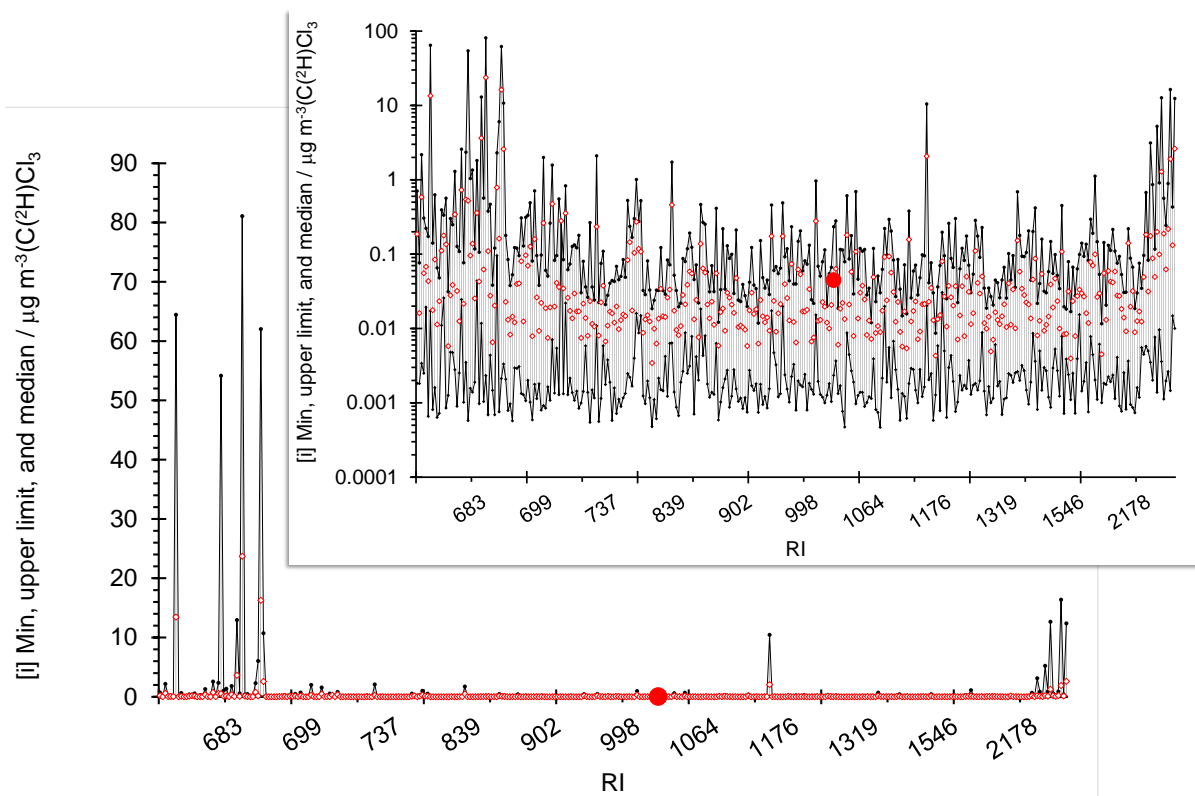
249 series data did not reveal any seasonal pattern, see Figure 5. Widening such an assessment
 250 to all of the 328 most frequently observed VOC revealed that most were present at
 251 concentrations below a threshold of $3 \mu\text{g m}^{-3}$ (expressed as a $\text{C}(\text{H})\text{Cl}_3$ equivalent) with 14
 252 compounds present at higher concentrations in the range 10 to $100 \mu\text{g m}^{-3}$, see Figure 6.
 253 The compounds present at the highest concentrations appear to be associated with solvent
 254 and disinfection formulations (2-propanol for example), or, higher molecular weight waxes
 255 associated with the use of nitrile protective gloves. The more volatile the contaminant the
 256 more frequently it was observed.



257

258 Figure 5 A summary of the environmental concentrations of octanal observed over the
 259 study. Top left, the distribution of octanal concentration; top right, the time series
 260 concentration data of background octanal concentration; and, bottom the
 261 octanal levels indicated by circles with the dashed line showing the upper limit
 262 from the distribution in the top left figure. The bottom figure's whiskers indicate
 263 the 20-fold threshold widely applied for the inclusion/exclusion threshold for
 264 admitting a VOC into a breath biomarker discovery and validation data set. The
 265 non-reproducibility and variability of such an approach may be discerned.

266



267

268 Figure 6 The minimum, median (red diamonds) and upper limit of observed
 269 concentrations across the 328 most frequently observed (> 5%) VOC over the
 270 duration of the study. The octanal entry is denoted by the solid red circle. The
 271 concentrations of these compounds spanned a range of 2 to 3 orders of
 272 magnitude and were typically below $3 \mu\text{g m}^{-3}$ (expressed as a $\text{C}(\text{H})\text{Cl}_3$ equivalent)
 273 with the exception of volatile solvents (RI less than 690) and heavier waxes
 274 associated with PPE production (RI greater than 2180).

275 **DISCUSSION**

276 This study describes the complexity of the contamination profiles and concentration
 277 distributions of VOC in the indoor air of hospitals and highlights seasonal and clinical
 278 variations. Ninety five percent of the compounds observed occurred in in less than 5 % of
 279 samples, creating a highly variable and non-reproducible VOC profile. Such variability may
 280 be attributed to the constantly changing demographic of the occupants of a busy clinical
 281 facility, combined with the range of therapies and therapeutics being administered. A less
 282 variable constituency of the background contained a mixture of ubiquitous exogenous VOC.
 283 Further, evidence that exhaled VOC may also be considered as VOC source in a clinical
 284 environment, and that seasonal factors were also present, was noted.

285 Exhaled VOC

286 Changes in the VOC composition of room air from exhaled breath have been described with
287 on-line proton transfer mass spectrometric studies reporting the presence of low molecular
288 weight exhaled VOC, with estimated exemplar office air concentrations for acetone, ethanol,
289 and isoprene of 52, 32, and 13 $\mu\text{g m}^{-3}$ respectively³⁸, while studies with cinema audiences
290 have monitored, and coded, VOC levels in room air to the emotional states of the scenes
291 being viewed³⁹. Other VOC associated with skin volatiles and the gastric tract were also
292 reported as constituents of the cinema auditorium's air. At a significantly larger scale,
293 changes in VOC levels in stadium air have been reported at sports events with isoprene,
294 acetone and ethanol increasing to 8.5, 9.7 and 580 $\mu\text{g m}^{-3}$ respectively during a football
295 match⁴⁰. Such studies suggested that the observation of exhaled volatiles in clinical room
296 air was to be anticipated, and that the 10^4 sample enrichment obtained through two stage
297 thermal desorption, combined with temperature programming, would enable concentrations
298 lower than 1 $\mu\text{g m}^{-3}$ to be monitored.

299 Exhaled VOC at trace levels were found in the room air of the clinics studied, most notably
300 the two metabolites of sevoflurane, for it is difficult to conceive of an alternative source. The
301 aldehydes in Figure 4, noted as oxidative stress biomarkers associated with respiratory
302 disease, have also been associated with aging and outgassing from linoleum flooring¹⁶, and
303 decanal is also an oxidative product of skin lipids. It seems plausible to propose that the
304 observed airborne concentrations of aldehydes were derived from a combination of constant
305 background emissions due to outgassing from flooring and building materials, overlaid with
306 concentration transients from exhaled breath combined with changes in room occupancy
307 and associated ventilation.

308 Seasonal VOC.

309 Previous studies on VOC domestic air contaminants¹⁶ have considered seasonal factors in
310 indoor contamination profiles and levels. Factors such as seasonal changes in ventilation

311 and indoor-based activities were identified. Further, the effect of solar-radiation on building
312 materials and outgassing was highlighted. The VOC in Table S2 that differentiate between
313 the summer and winter months may be attributed to sources that include: cleaning and
314 disinfection materials (propan-2-ol, (1R,2S,5R)-2-isopropyl-5-methylcyclohexanol
315 [dimenthol], (6S)-2,6-Dimethyl-7-octen-2-ol, (6S)-2,6-Dimethyl-7-octen-2-ol and 2-
316 butoxyethanol); outgassing from plasticiser/polymer components (2-ethyl-1-hexanol and
317 2,5-Cyclohexadiene-1,4-dione-2,5-diphenyl, phenol, 1-methylethylbenzene
318 [isopropylbenzene] and cyclohexanone); pollutants associated with fugitive emissions from
319 vehicles and their exhausts (toluene, ethylbenzene, benzene, 1,4-dimethylbenzene and 1,2-
320 dimethylbenzene, 1-methylethylbenzene); and exhaled volatiles (cyclohexanone, ethanol
321 and ethanal). Such seasonal changes are consistent with changes in ventilation, noted
322 previously¹⁶, and increased usage of cleaning products associated with higher bed
323 occupancy rates during the winter season. Increased concentrations of ethanol and its
324 catabolite, ethanal are consistent with winter patterns of alcohol consumption, combined
325 with changes in ventilation.

326 Breath-biochemistry research and biomarker discovery and monitoring studies manage
327 room air contamination compounds in different ways. Some studies invoke an “alveolar
328 gradient” concept⁴¹ and others remove compounds from the data-processing workflow
329 where the environmental background is greater or equal to 5% of the exhaled concentration.
330 An alveolar gradient approach is problematic because: the data presented in this paper
331 indicate that important disease markers are routinely also present in room air; and in a
332 clinical environment a participant’s exposure to a specific VOC is unknown, as is the
333 subsequent rate of the compound’s elimination prior to the breath sample being taken.
334 Another approach that excludes breath data if it is detected in the environment⁴² is also
335 unsatisfactory for the first reason given above.

336 Alveolar gradient and environmental exclusion approaches are unlikely to resolve the
337 confounding factor of variable environmental backgrounds, participant exposures, and
338 participants' catabolisms. Further, such approaches also create the possibility of
339 uncontrolled, and variable, site and time dependent factors determining whether or not
340 participant data is admitted into a study. Figure 5 shows a range of inclusion concentrations
341 (20 times the environmental background that might be applied over the 245 breath data
342 sets acquired) for octanal with a range from ca. 20 ng m⁻³ to 2 µg m⁻³ (expressed as a
343 C(2H)Cl₃ equivalent). Adopting a concentration gradient approach means that the inclusion
344 criteria for breath data varies between participants in an apparently arbitrary manner with
345 little or no evidence to relate the studies threshold concentration (for that particular sample)
346 to exposure or washout.

347 With the exception of a few solvents (formulation compounds) and some mould release
348 agents, associated with disposable PPE, most volatiles, when present, were below a
349 threshold, see Figure 6. In the case of octanal this was 82 ng m⁻³ (expressed as a C(2H)Cl₃
350 equivalent), and an exhaled breath concentration above this level was highly unlikely to be
351 due to environmental exposure. A study of the 328 common pollutants indicates that a
352 default level of 3 µg m⁻³ (expressed as a C(2H)Cl₃ equivalent) is a threshold concentration
353 above which most clinical room air VOC do not rise. In the absence of reliable site-specific
354 environmental survey data applying such a threshold is a more reliable, reproducible and
355 systematic approach than adopting the variable levels acquired by spot sampling. Infrequent
356 VOC contamination transients would be identified as outliers and rare events in the exhaled
357 breath profiles and be excluded from the data pipeline in the event they occurred.

358 Seasonal factors mean that studies and clinical breath testing need to account for possible
359 bias. More robust discovery designs will randomise recruitment of participants over the
360 seasons of the year and avoid overweighting those months with higher occupancy and

361 higher clinical VOC levels. Such attention to scheduling might also be usefully applied to
362 sampling sessions in clinic to identify other possible synchronisation of environmental
363 exposure with sampling activities; cleaning, cooking and medication for example.

364 VOC that are frequently present within a facility at elevated concentrations may introduce a
365 “contrast agent” into breath studies, in that differentiation between classes of participants
366 may be based on differences in how the contaminant is metabolised, arising from catabolic
367 changes due to disease or treatment. Such a situation may result in fortuitous discovery, or
368 non-reproducible results. It will be helpful to note and report background contaminants
369 routinely detected at elevated concentration (i.e., above $3 \mu\text{g m}^{-3}$) and whether they may be
370 modelled as part of breath data or not.

371 Finally, adopting a concentration threshold approach does not remove the necessity for
372 obtaining reliable environmental background data as part of breath testing operations, such
373 samples and analysis should still an integral part of a quality assurance and control protocol.

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- 554

555 SUPPLEMENTARY DATA

556

557 THE VARIABILITY OF VOLATILE ORGANIC COMPOUNDS IN CLINICAL ENVIRONMENTS.

558 Dahlia Salmana, Wadah Ibrahim^{b,c}, Amisha Kanabara, Bo Zhao^{b,c}, Amisha Singapuri^{b,c}, Michael
559 Wilde^d, Rebecca L Cordell^d, Teresa McNally^b, Dorota Ruszkiewicz^a, Andria Hadjitheklia, Robert Free^{b,c},
560 Neil Greening^{b,c}, Erol A Gaillard^b, Caroline Beardsmore^b, Paul Monks^d, Chris Brightling^{b,c}, Salman
561 Siddiqui^{b,c}, C L Paul Thomas^{a*} On behalf of the EMBER consortium

562 ***Corresponding author:** C.L.P.Thomas@lboro.ac.uk

563 a Department of Chemistry, Loughborough University, Loughborough, LE11 3TU, UK.

564 b College of Life Sciences, Department of Respiratory Sciences, University of Leicester,
565 University Road, Leicester LE1 7RH

566 c Leicester NIHR Biomedical Research Centre (Respiratory theme), Glenfield Hospital, Groby
567 Road, Leicester LE3 9QP

568 d Department of Chemistry, University of Leicester, University Road, Leicester, LE1 7RH

569

570

571 FIGURES

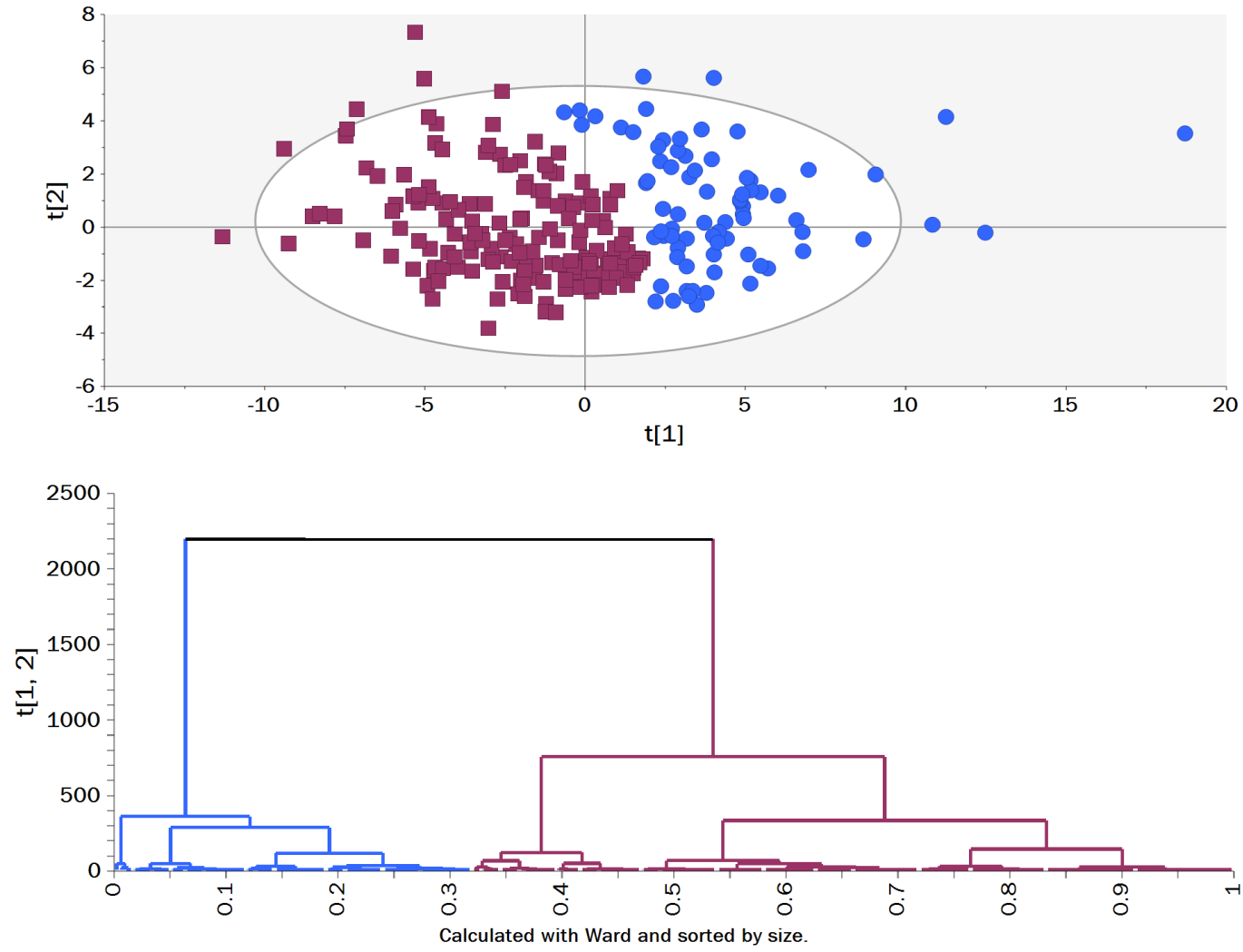
572



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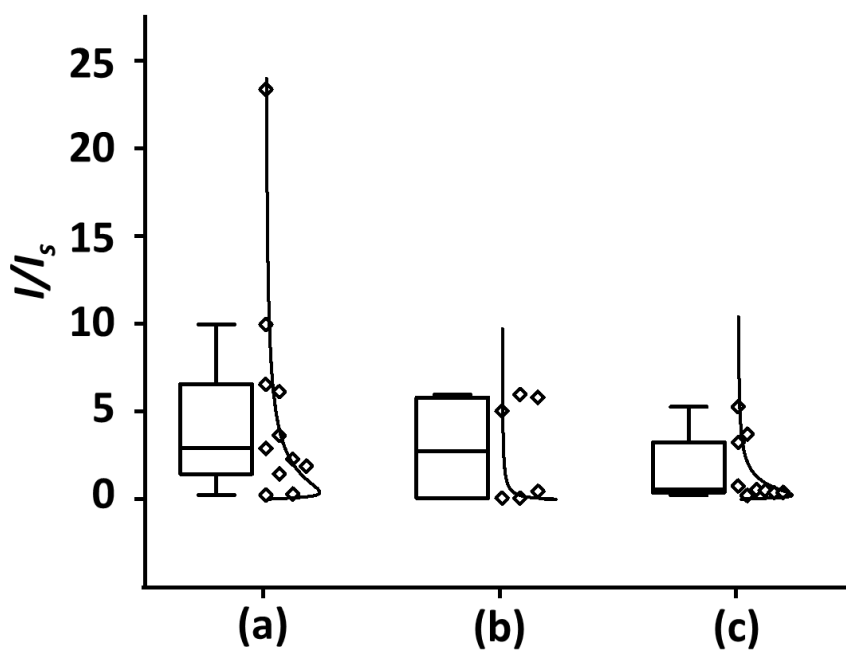
574 Figure S7: The Escort Elf pump attached to a Tenax®/Carbotrap 1TD hydrophobic adsorbent
575 tube to continuously sample (500 cm³ min⁻¹) the indoor air of the clinic/ward
576 for 120 s.

577



578

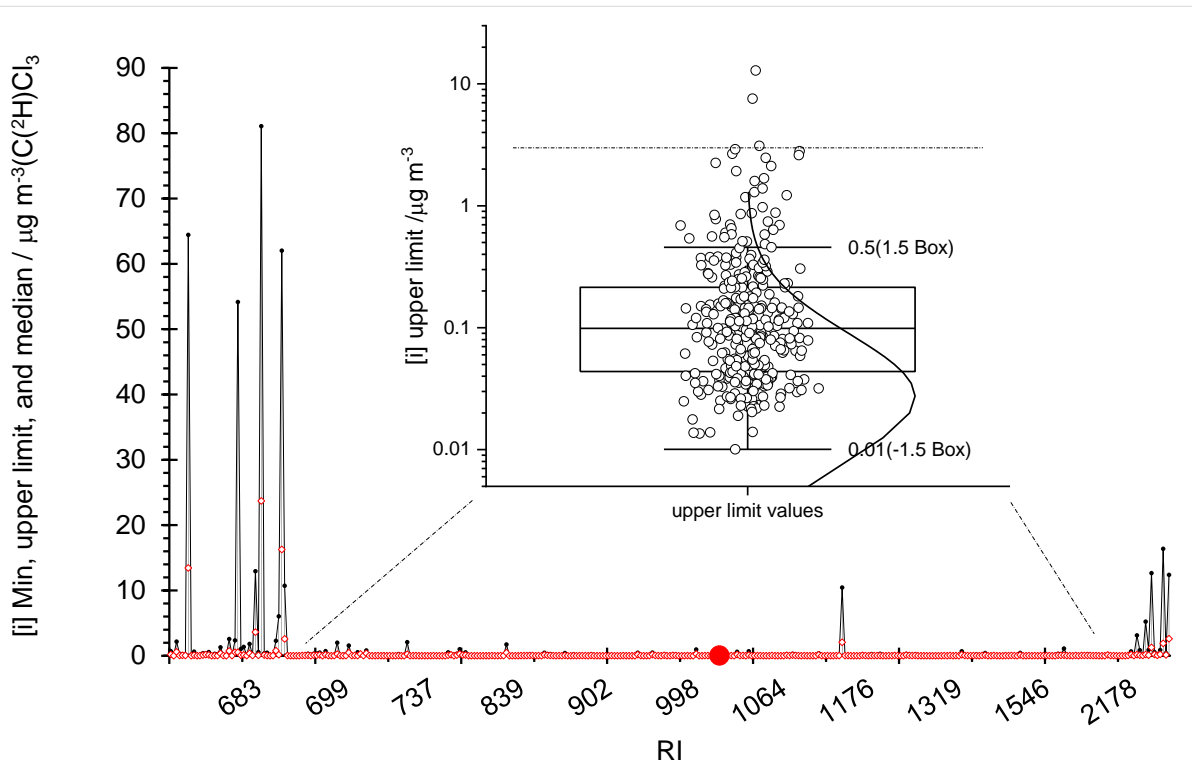
579 **Figure S8:** PCA (top) and its associate dendrogram (bottom) to examine the seasonality effects on the environmental VOCs profile in the
 580 two clinics. Winter and summer groups are coloured with blue and red, respectively.



581

582 **Figure S9** Box plots and lognormal distribution curves for (a) Sevoflurane (m/z 131,
 583 retention index 661, mean 3.09, sd 5.51) and its two metabolites (b) 2-
 584 Propanone, 1,1,1,3,3,3-hexafluoro (m/z 69, retention index 655, mean 0.92,
 585 sd 2.04) and (c) and 2-Propanol, 1,1,1,3,3,3-hexafluoro (m/z 51, retention
 586 index 670, mean 0.79, sd 1.48)

587



588

589

Figure S10

Further detail on Figure 6 comparing the minimum, median (red diamonds) and upper limit of observed concentrations across the 328 most frequently observed (frequency > 5%) VOC over the duration of the study. The octanal entry is denoted by the solid red circle. The concentrations of these compounds spanned a range of 2 to 3 orders of magnitude and the Q3+1.5 IQR was 0.5) The inset box whisker plot consolidates the upper limits of the 328 VOC. A working threshold of $3 \mu\text{g m}^{-3}$ (expressed as a $\text{C}(\text{H})\text{Cl}_3$ equivalent and indicated by the dotted line) may be reliably applied across the RI range of RI 700 to RI 1900. This excludes volatile solvents (RI less than 700) and heavier waxes associated with PPE production (RI greater than 1900),

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602 TABLES

603 Table S2: TD-GC-MS operational conditions

Thermal desorption		Gas chromatography		Mass spectrometer	
Parameters	Setting	Parameters	Setting	Parameters	Setting
<i>t</i> Primary desorption	1 min	<i>F</i> He carrier gas	2 cm ³ min ⁻¹	Scan type	Full scan (+ve)
<i>F</i> Primary desorption	40 cm ³ min ⁻¹	<i>T</i> Initial	40 °C	Mass range	40 to 550 m/z
<i>T</i> Primary desorption	300 °C	<i>t</i> Initial hold	0 min	Ionisation type	EI
<i>T</i> Secondary desorption	5 min	<i>T</i> program	5 °C min ⁻¹ to 300 °C	<i>v</i> scan	3 Hz
<i>F</i> Secondary desorption	50 cm ³ min ⁻¹	<i>T</i> End	300 °C	<i>T</i> line temperature	300 °C
<i>T</i> Secondary desorption	300 °C	<i>t</i> End hold	0 min	<i>T</i> Quadrupole	150 °C
<i>F</i> Cold trap	20 cm ³ min ⁻¹	<i>t</i> Total run	60 min	<i>T</i> Manifold	230 °C
<i>T</i> Cold trap	-10 °C	<i>T</i> Post run	45 °C	<i>t</i> Solvent delay	5 min
$\left(\frac{\delta T}{\delta t}\right)_{\text{trap}}$	Max °C min ⁻¹	<i>t</i> Post run	0 min		
<i>T</i> Trap high	300 °C				
<i>t</i> Trap hold	5 min				
<i>T</i> Flow path	200 °C				
Mode	Spitless				

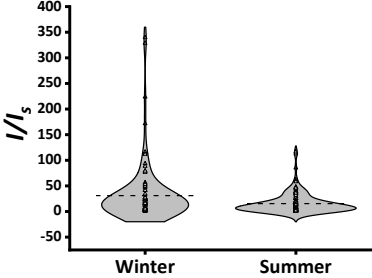
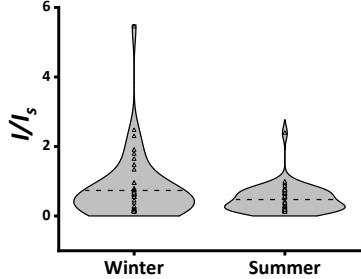
604 Note: **t**, time; **F**, flow; **T**, temperature; and **v**, frequency.

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606

607 **Table S3:** A list of volatile organic compounds that were detected in the environmental air of both hospitals and were found to be higher
 608 in the winter seasons than summer.* is for compounds that were higher in winter by 1.5-fold however not statistically different
 609 following a one-tail t-test. †Identification of ethanol is qualified in that the full mass spectrum was not captured within the scan
 610 range of the method.

611 \bar{I}_W/I_{IS} : mean intensity ratio to C(2H)Cl₃ internal standard in winter; S_W standard deviation (winter); \bar{I}_S/I_{IS} : mean intensity ratio
 612 to C(2H)Cl₃ internal standard in summer; S_S standard deviation (summer); \bar{I}_W/\bar{I}_S fold change in abundance between winter and
 613 summer; DoF : degrees of freedom

Compound	Winter		Summer		\bar{I}_W/\bar{I}_S	t Stat	t_{crit} 1-tail	$P \leq t$ 1-tail	DoF	Violin plots
	\bar{I}_W/I_{IS}	S_W	\bar{I}_S/I_{IS}	S_S						
Ethanol	30.98	55.89	15.44	19.06	2.01	2.63	1.65	0.004	113	
Cyclohexanone	0.73	0.90	0.47	0.38	1.55	1.79	1.67	0.038	62	

Benzaldehyde

1.54

3.65

0.69

0.51

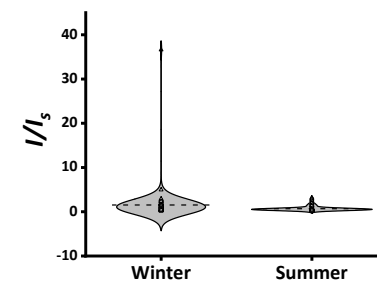
2.23

2.27

1.66

0.010

100



Propan-2-ol

149.49

304.00

68.82

100.02

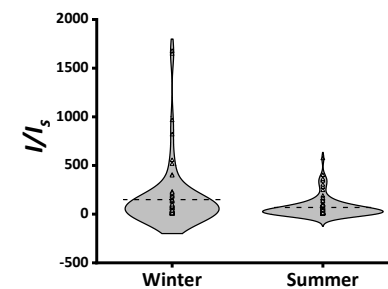
2.17

2.17

1.66

0.016

85



Toluene

2.49

3.79

1.07

0.93

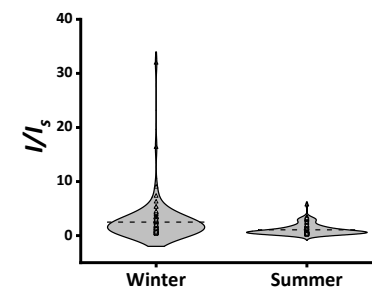
2.33

3.48

1.66

0.0004

99



Ethyl benzene

0.66

0.60

0.38

0.39

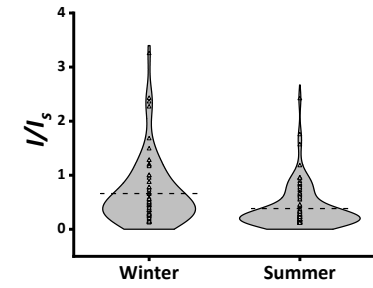
1.72

3.35

1.66

0.0006

111



Phenol

1.58

4.36

0.74

0.53

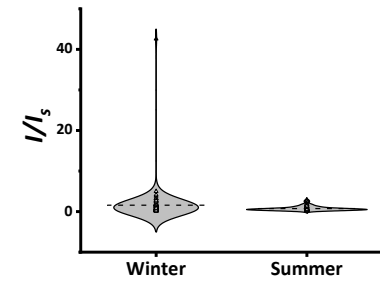
2.13

1.84

1.66

0.034

95



Benzene

2.23

4.00

1.40

0.68

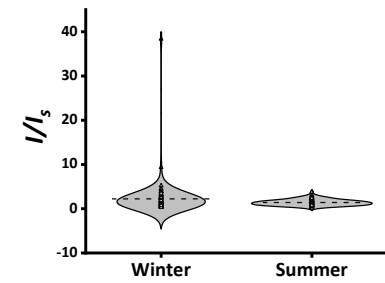
1.59

1.96

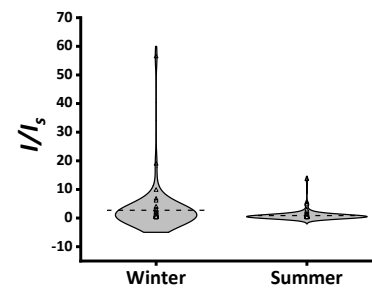
1.66

0.0260

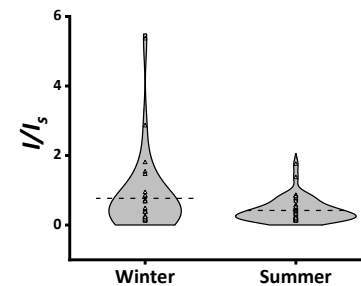
96



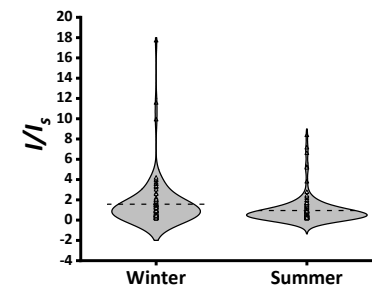
1,4-dimethylbenzene	2.73	7.93	0.88	1.65	3.11	1.67	1.67	0.0499	56
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2-butoxyethanol	0.77	1.04	0.42	0.32	1.83	1.76	1.69	0.0441	32
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2-ethyl 1-hexanol-	1.57	2.38	0.94	1.29	1.67	2.30	1.66	0.0113	133
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**(6S)-2,6-Dimethyl-
7-octen-2-ol**

1.11

2.61

0.47

0.48

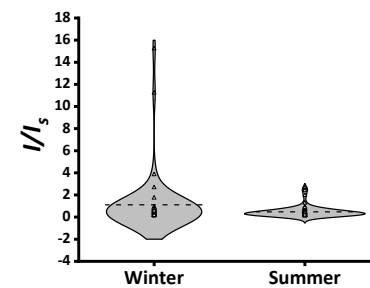
2.36

1.68

1.68

0.0491

51



Ethanal *†

9.36

22.37

4.88

20.75

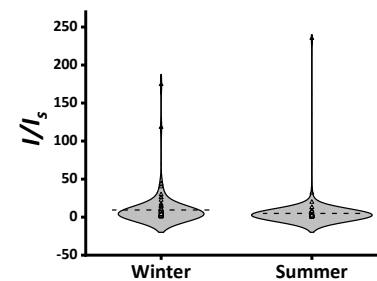
1.92

1.49

1.65

0.068

185



**2,5-
Cyclohexadiene-
1,4-dione, 2,5-
diphenyl- ***

4.65

9.47

2.16

8.17

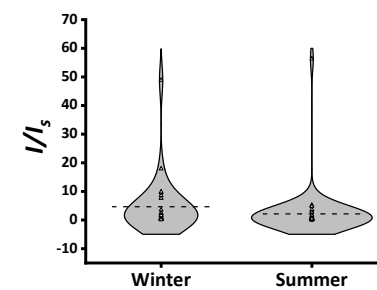
2.15

1.13

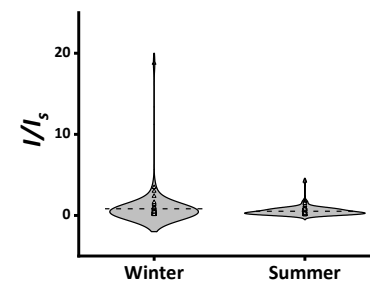
1.68

0.130

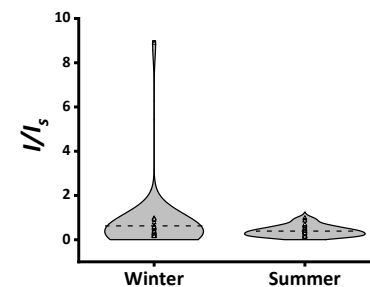
50



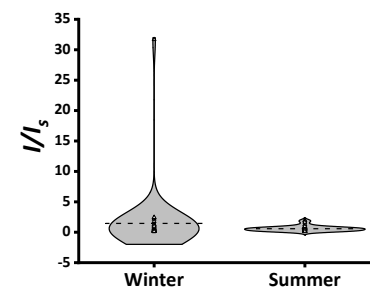
1,2-dimethylbenzene *	0.82	2.21	0.51	0.52	1.60	1.14	1.66	0.1280	77
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Benzene, (1-methylethyl)- *	0.63	1.42	0.39	0.23	1.61	0.97	1.68	0.1669	37
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(1R,2S,5R)-2-isopropyl-5-methylcyclohexanol *	1.45	5.01	0.59	0.42	2.44	1.04	1.69	0.1517	37
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