

Platinum Interdigitated Electrode Arrays for Reagent-free Detection of Copper.

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Received: date; Accepted: date; Published: date

Abstract: Water is a precious resource that is under threat from a number of pressures, including release of toxic compounds that can have devastating effect on ecology and human health. The current methods of water quality monitoring are based on sample collection and analysis at dedicated laboratories. This can provide an incomplete picture of the status of the water body as pollution events can be missed. Recently, electrochemical based methods have attracted a lot of attention for environmental sensing owing to their versatility, sensitivity and compatible integration with cost effective, smart and portable readout systems. In the present work, we report on the fabrication and characterization of platinum interdigitated microband electrodes arrays, and their application for trace detection of copper. Using square wave voltammetry after acidification with mineral acids, a limit of detection of 0.8 $\mu\text{g/L}$ was achieved. Copper detection was also undertaken on river water samples and compared with standard analytical techniques. The possibility of controlling the pH at the surface of the sensors – thereby avoiding the necessity to add mineral acids – was investigated. By applying potentials driving the water splitting reaction at one comb on the sensors electrode (the protonator), it was possible to lower the pH in the vicinity of the sensing electrode. Detection of standard copper solutions down to 5 $\mu\text{g/L}$ using this technique is reported. This reagent free method of detection opens the way for autonomous, in situ monitoring of pollutants in water bodies.

Keywords: Electrochemical Sensors; Environmental Monitoring, Heavy Metals, pH Control

1. Introduction

An essential component in human development and survival is access to clean and safe water [1]. Unfortunately, water can be contaminated by a variety of pollutants such as pesticides used in agriculture, antibiotics from both human and animal consumption or heavy metals from manufacturing. Pollution from heavy metals is a serious concern due to their adverse effects caused to human health. [2] Of these heavy metals, copper is widely distributed in the environment due to its use in a range of anthropogenic processes. Copper can be found in many fertilizers in agriculture [3]. It is also an essential component in the manufacturing of electronics [4] and improper disposal of electronic products can lead to excess copper in the environment. Many pipes in old plumbing systems in Ireland use copper piping. This can lead to copper leaching into the water system [5]. Finally, mining activities or mismanagement of abandoned mines can also be another source of copper run-off [4].

Excess heavy metal pollution in the environment presents a major threat for many ecosystems. It can affect aquatic systems and animals [3]. Because these metal ions are persistent and do not break down further in environmental systems [6], they can then bioaccumulate through the food chain, eventually affecting human food produce [4]. While trace amounts of some heavy metals are essential for life [7], these can be detrimental to human health at higher concentrations. This includes copper [8]. For this reason, the World Health Organization (WHO) and the European Union (EU) have set the maximum permissible allowed concentration of copper in drinking water as 2 mgL^{-1} [9]. Increased levels of

copper can lead to various human disorders such as Wilson's Disease [10, 11], Alzheimer's [12], Parkinson's [13], Prion, Huntington's disease [14], and disorders such as diabetes and obesity [15]. Other symptoms caused by excess copper include stomach cramps, vomiting and gastrointestinal issues [16]. As a result, environmental monitoring of copper levels is of critical importance for both ecological and public health preservation. Ideally, a monitoring system capable of in situ, real time and sensitive detection could provide early warning pollution alert that could possibly be used to mitigate the impact of pollution events.

Conventional analytical detection methods for heavy metals such as copper include Atomic Absorption Spectroscopy (AAS) [17], Inductively Coupled Plasma Mass Spectrometry (ICP-MS) [18] and Inductively Coupled Plasma Atomic Emission Spectrometry (ICP-AES) [19]. Fluorescent [20], colorimetric [21] and spectroscopic methods such as Raman [22] and Surface Enhanced Raman Scattering [23] have also been used for detection of heavy metals. However, these methods are laboratory-based and require expensive, bulky instrumentation, trained personnel and some require sample pre-treatment making them unsuitable for on-site or in situ testing. In this respect, electrochemical methods have attracted a lot of attention since they can provide a rapid, cost effective, and sensitive method for on-site analysis of heavy metals without the need for lab or trained operators. Furthermore, the readout electronic boards can host data analytics protocols and wireless communication technologies enabling smart, autonomous, networked, remote environmental monitoring system [24].

A variety of electrode materials have been employed for detection for copper, see Table 1. These including glassy carbon [3], imprinted polymeric film [8], zeolite/expanded graphite structures [19], multi-wall carbon nanotubes-carbon paste [25], screen printed electrodes [2, 26], sol-gel Au nanoparticle-carbon paste electrodes [27], quantum dots [28, 29], graphene like carbon [28, 30], iridium micro disks [31], surfaces modified with chelating agents [16, 32], and metal oxides [28]. Current electrochemical methods are mainly based on macro sized electrodes. However, sensors based on microscale electrodes have been shown to have enhanced mass transport properties, leading to faster response time with improved sensitivities.

Table 1 Electrochemical sensors for copper detection

| Electrode type | Linear Range | LOD | Ref. |
|---|---|-------------------------|------|
| Glassy Carbon | 0.5 - 80 mM | 0.44 mM | [1] |
| Imprinted Polymeric Film | 0.9-150 nM | 2.7 nM | [2] |
| Zeolite/Expanded Graphite Structures | | 2.23 nM | [3] |
| Multi-wall Carbon Nanotubes-Carbon Paste | 1 mM-0.1 M | 0.79 mM | [4] |
| Screen Printed Electrodes | 10 mM-150 mM | 1.4 nM | [5] |
| Sol-gel Au Nanoparticle-Carbon Paste Electrodes | 0.43 mM-10 mM | 0.4 mM | [6] |
| Quantum Dots | 10-1100 nM | 6 nM | [7] |
| Graphene like Carbon | | 0.23 nM | [8] |
| Iridium Micro Disks | 20 - 100 ppb | 5 ppb | [9] |
| Surfaces Modified with Chelating Agents | 1.0×10^{-9} - 1.0×10^{-4} M | 8.91×10^{-9} M | [10] |
| Metal Oxides | 2 - 100 nM | 1.08 nM | [11] |

A variety of electrochemical techniques can be used for sensing applications. Amongst them, square wave anodic stripping voltammetry (SWV) has been shown to be particularly suitable for detection of

heavy metals. This method involves two steps, the first step being the pre-concentration of the target metal at the surface of the electrode. This step is then followed by oxidative stripping of the metal from the surface by sweeping the potential through the metal's oxidation potential. Detection limits of copper in the $\mu\text{g/L}$ range can be readily achieved using this method.

An important consideration for the electrochemical detection of heavy metals is the sample pH. Indeed, the pH can affect the availability of metal ions for electrochemical processes. Generally, electrochemical methods are carried out with addition of acids in order to chemically adjust the pH to that needed for the measurements. This means that some sample preparation is required for the analysis. This cannot be easily integrated into portable devices, meaning true in situ measurements are impossible.

The approach investigated here is to negate the need for chemical reagent addition completely by incorporating on chip local pH control. This is achieved by the electrolytic decomposition of water to form H^+ ions near the working electrode. The sensors are comprised of two interdigitated arrays of platinum (Pt) electrodes. Electrode microbands are spaced $2\ \mu\text{m}$ apart. One side, or comb, of the array is used as the working electrode and carries out the SWASV as it would normally. The opposite side of the comb is held at a constant potential to adjust the pH close to the electrode to the required pH. This method has been shown to work for the detection of free chlorine [35], silver [36] and has also been used for the elimination of oxygen interference in the detection of monochloramine [37].

2. Materials and Methods

2.1 Chemicals and Solutions:

Copper sulphate (CuSO_4), Nitric acid (HNO_3), Sodium Chloride (NaCl), Phosphate Buffer Saline (PBS) and ferrocenecarboxylic acid (FCA) were obtained from Sigma Aldrich Ireland. All aqueous solutions were made up using ultra-pure Milli-Q water ($18.2\ \text{M}\Omega\cdot\text{cm}$, Milli-Q). Standard solutions of copper were made up using $10\ \text{mM}$ NaCl and diluted as needed from a mother solution. These standards were adjusted to pH 2 by addition of $1\ \text{M}$ HNO_3 added dropwise until the desired pH was achieved. All chemicals were used as received without any further purification.

2.2 Instrumentation:

All electrochemical measurements were undertaken using a portable CH Instrument 1220C bi-potentiostat. A three-electrode configuration was used for copper detection using platinum interdigitated microband arrays (IDA) as the working (or sensing) electrode, and gold and platinum on-chip counter electrode and pseudo reference electrode, respectively. In some instances, an Ag/AgCl reference electrode was used – using this electrode was found to shift the cyclic voltammogram by *ca* $+250\ \text{mV}$ and all potentials (deposition potential, acquisition potential window) were shifted accordingly. An additional platinum interdigitated microband array (the protonator) was introduced and used in a four-electrode configuration for electrochemical pH adjustment.

2.3 Sensor Design and Fabrication

The fabricated sensor chips were $20\ \text{mm}$ by $8.5\ \text{mm}$ with 19 connection pinouts on the left-hand side. The design was compatible with HDMI-C connector, permitting facile electrical connection to external electronics. Sensor chip fabrication was similar to the one described by Wahl et al. [38] and was based on lithographic processes commonly used in microelectronic foundries, and are therefore compatible with mass manufacturing. The sensors were fabricated on four-inch silicon wafer substrates with a $300\ \text{nm}$ layer of thermally grown silicon dioxide. Firstly, the working electrodes were patterned using photolithography and thermal evaporation ($50\ \text{nm}$ of Pt, with $10\ \text{nm}$ of Ti adhesion layer) followed by

lift off. A second optical lithographic and metal deposition process (Ti 10 nm/Au 100 nm) was undertaken to define HDMI pin-out, interconnection tracks, as well as the on-chip counter electrode (500 μm wide x 6 mm long). A third similar step was used to define the Pt (Ti 10 nm/Pt 100 nm) on-chip reference electrode (500 μm wide x 6 mm long). Finally, 500 nm of PECVD SiN was blanket deposited on the whole wafer, and openings over the working/counter/reference electrodes and electrical contacts defined by lithography and dry etching. An optical image of a fabricated chip can be seen in Figure 1 (a). A die contained 8 individual sensors, allowing for multiplexed measurements if needed. Each sensor comprised two interdigitated arrays (IDA); the protonator IDA comprised 14 tines while the working IDA comprised 13 tines. Each tines were 1 micron wide and 45 microns long, and the gap between tines was 2 microns, see high magnification in Figure 1 (a) and (b). A HDMI-C pinout was also incorporated on one side of the chip to allow facile electrical connection to the electrodes. Prior to the experiments, each chip was inspected using optical microscopy to identify any obvious defects or faults. The chips were washed with isopropanol and deionised water to remove any residual contaminants from the electrode surface. The overall experimental setup comprised the sensors chips, the HDMI-C connector, a portable bipotentiostat and a laptop, see Figure 1 (c), and could easily be brought on the field for on-site measurements.

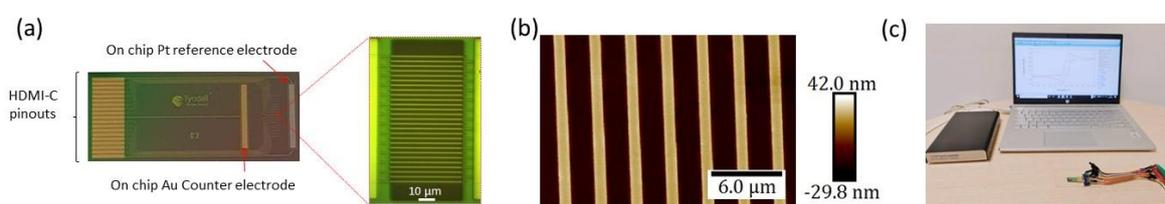


Figure 1 (a) optical micrograph of the HDMI-C sensor chip, showing the Au counter electrode, the 8 Pt working electrodes and the Pt reference electrode ; with higher magnification of one of the working electrode. (b) Atomic Force microscope image of the working electrode. The interdigitated microbands are 1 micron wide with 2 micron gaps (c) Experimental setup comprising sensor chip, connector, portable bi-potentiostat and laptop.

2.4 Optical and surface morphology characterization

Optical micrographs were acquired using a calibrated microscope (Axioskop II, Carl Zeiss Ltd.) equipped with a charge-coupled detector camera (CCD; DEI-750, Optronics). The surface morphology of the sensors was studied using a Bruker Nanoscope dimension icon atomic force microscope in tapping mode.

2.5 Copper Detection using Chemical pH adjustment

The electrochemical detection of copper was carried out in a multi-step process, see Figure 2. First, a 50 μL aliquot of the solution to analyse, adjusted to pH 2 with nitric acid, was pipetted onto the active area of the sensor chip. The next step was the electrodeposition of copper ions present in solution onto the surface of the platinum working electrode. This was achieved by applying a potential of -0.65V to the Pt working electrode. Following this, square wave voltammetry from -0.7 to -0.2V was used to strip the as deposited copper back into the solution. This resulted in an oxidation peak with a peak height proportional to the concentration of copper present in solution.

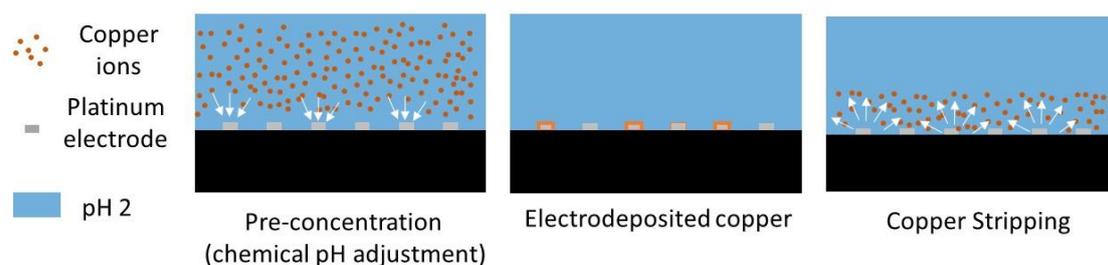


Figure 2 Schematic of electrochemical detection of copper. Step 1: pre-concentration at the surface of the working electrode; step 2: stripping of deposited copper back in solution

2.6 River water sample Collection

Analysis of water collected at Avoca (Co. Wicklow), Ross island (Co. Kerry) and Bunmahon (Co. Waterford), was undertaken, see Figure 3 (a). These sites were selected as they were expected to have high copper levels due to their proximity to disused copper mines [39], and could also be easily and safely accessed. Water samples were analysed on-site (i.e. by taking a sample of water, and carrying out the analysis on the side of the water body), see Figure 3 (b). Grab samples were also brought back to the laboratory for further analysis and sent for ICP-MS analysis by an accredited laboratory (Environmental Laboratory Services, Cork).



Figure 3 (a) Map of Ireland showing the sampling/measurement locations, (b) measurement being carried on-site at Avoca river, co Wicklow.

3. Results and Discussion

3.1 Electrode Array Characterisation

Cyclic Voltammetry (CV) was carried out in 1 mM FCA in 10 mM PBS in a potential range of -0.2 V to 0.4 V to characterize the electrochemical functionality of the sensors. As can be seen in Figure 4 (a), the CV has a typical duck shape indicative of diffusion limited behavior commonly observed at macroelectrodes. This arose as a result of the overlapping of radial diffusion profiles surrounding individual microband of an IDA, thus resulting in an overall time-dependent diffusion limited behaviour [40]. The interdigitated setup was also tested with FCA in a collector generator mode, see Figure 4 (b). In this approach, one comb of the electrode (the generator) was cycled as above while the other comb (the collector) was held at -0.1 V. As in the previous case, the FCA molecules begin to oxidize to FCA⁺ at the generator for potentials above 0 V. The oxidized species then diffused above the collector where they were reduced back to FCA. As a result of this process, the CV exhibited a steady state profile with higher measured currents (40 nA vs 9 nA). The collection efficiency of the sensor, defined as the ratio of the collector to the generator currents, was determined to be ~84 %.

This suggests that for electrochemical pH adjustment (*vide infra*), 84 % of protons generated at the protonator, would diffuse to the working IDA.

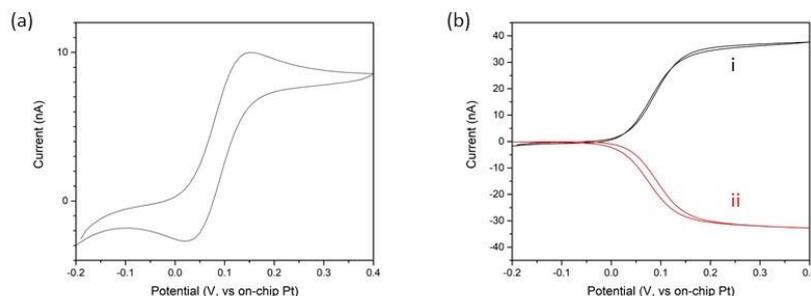


Figure 4 (a) Typical cyclic voltammogram in 1 mM FCA in 10 mM PBS measured at a platinum interdigitated electrode comb; (b) CVs at the generator and collector IDAs. The generator IDA (i) was cycled between -0.2 V and 0.4 V while the collector IDA (ii) was held at -0.1V.

3.2 Detection of Cu in standard solutions using chemical pH adjustment

Figure 5 (a) shows the cyclic voltammogram for copper deposition from a 10mM aqueous solution of CuSO_4 at pH 2. An anodic current corresponding to the reduction of copper ion to copper metal could be seen starting at a potential of approximately -0.65 V and continuing at lower potential. On the reverse sweep, the deposition continued up to -0.55 V, indicating that the growth of copper is more kinetically favoured on existing Cu nucleation sites. The return sweep also showed a sharp peak at -0.42 V corresponding to the stripping of the copper that was deposited. For the following experiments, a deposition potential of -0.65 V was applied to the working electrode for 300 s (-0.4 V when Ag/AgCl reference electrode was used). Using lower deposition potentials could result in higher amount of copper being deposited but, at low concentration, could also potentially increase the interference from hydrogen-evolution effects – a process particularly efficient at Pt electrodes. The deposited copper was then stripped from the surface using square wave voltammetry resulting in a peak at approximately -0.4 V. The following SWV conditions were used: starting potential -0.7V, frequency 25 Hz, increment potential 0.004 V, amplitude 0.05 V. A second SWV was carried out right after the first one to ensure that all the copper had been removed from the surface of the working electrode.

The influence of the solution pH on the peak stripping current response was studied by varying solution pH from 5 to 1 (100 $\mu\text{g/L}$ CuSO_4 solution, deposition at -0.65V for 300 seconds), see Figure 5 (b). Decreasing the pH led to a change in the measured stripping current with the maximum stripping peak current found at pH 2. This optimal pH value was thus selected as the pH of choice for further experiments.

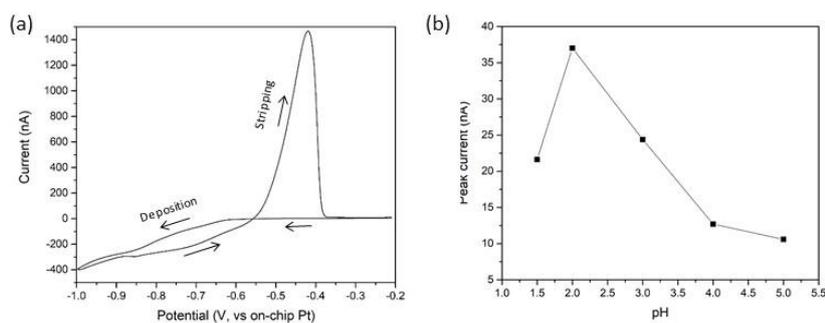


Figure 5 (a) Cyclic Voltammogram of copper electrodeposition onto the Pt microbands (10 mM CuSO₄, pH 2), (b) Influence of the pH of the solution on measured stripping current (100 µg/L CuSO₄ solution, 300s deposition at -0,65V). The line is a guide to the eye only.

Square wave stripping voltammetry (see section 2.5) was used to measure the levels of copper of serial dilutions of copper sulfate solutions adjusted to pH 2. Figure 6 (a) shows the stripping square wave voltammograms obtained for Cu concentrations ranging from 5 to 100 µg/L. The peak at approximately -0.25 V corresponds to the copper metal being oxidised and stripped from the surface of the electrode. As expected, the intensity of the peaks increased with increasing concentrations of copper in the solution. Concentrations down to 5 µg/L could be readily detected with the experimental conditions chosen. Figure 6 (b) shows the peak intensity as a function of Cu concentration. From this calibration plot, the linear equation $i(\text{nA}) = 0.85 + 0.37 C_{\text{Cu}}(\mu\text{g/L})$ with the correlation coefficient of 0.96 was obtained. The limit of detection (LOD) was calculated using equation 1:

$$\text{LOD} = 3 \times \text{SD}/S \quad (1)$$

Where SD is the standard deviation of the blank (measured to be 0.1 nA) and S the sensitivity of the electrode, defined as the slope of the calibration line [40]. A LOD of 0.8 µg/L was obtained. It should be noted that lower limit of detections could be achieved with longer deposition times or by enhancing mass transport with a microfluidic setup, but this was not investigated further in the present work as the reported LOD was well below the maximum permissible allowed concentration required by legislation. Also, the technique used for benchmarking was limited to 3 µg/L (*vide infra*).

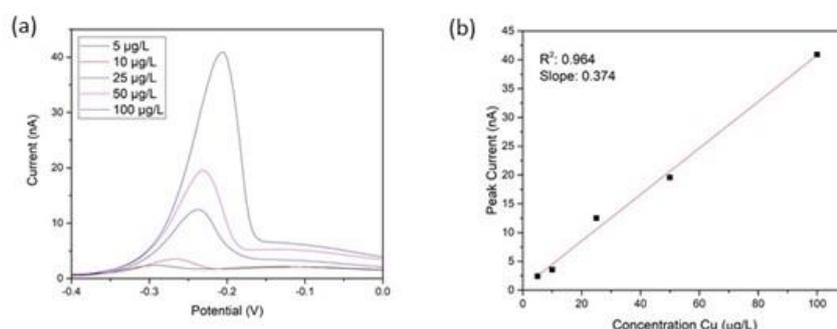


Figure 6 (a) Square Wave Voltammograms of Cu solutions at different concentrations (pH adjusted to pH 2 with acids, deposition for 300 sec at -0.4 V; (b) Corresponding calibration line.

3.3 Detection of Cu in real samples using chemical pH adjustment

In order to test our sensors with more complex matrixes, water samples were collected from sites where measurable levels of copper were expected (see section 2.6). Samples were grabbed from the water body and acidified to pH 2 on-site using nitric acid. While the collected samples did not show significant turbidity, they were rested for 5 minutes to allow bigger suspended particulates to fall at the bottom of the tube. Only water from the top of the test tube was used. Figure 7 below shows the stripping square wave voltammograms obtained after 300 s deposition at -0.65 V for the different sites. The measured peak current (peak to baseline) were 7.7, 9.3 and 1.1 nA for Avoca, Ross mine and Bunmahon, respectively. Using the calibration line established previously, this corresponded to copper concentrations of 17, 20 and 1 ppb, respectively. These water samples were also sent for ICP-MS analysis in accredited laboratory and concentrations reported were 22, 27 and < 3 ppb, respectively, see Table 2. Some discrepancy exists between the values measured using the present electrochemical sensors and those obtained with the commercial ICP-MS analysis. These could be due

to the electrochemical method adopted here, where sensors detect dissolved free copper only and not copper ions that are chelated to natural organic ligands or copper in suspended solids form such as particulates or colloids - thus under reporting the total copper.

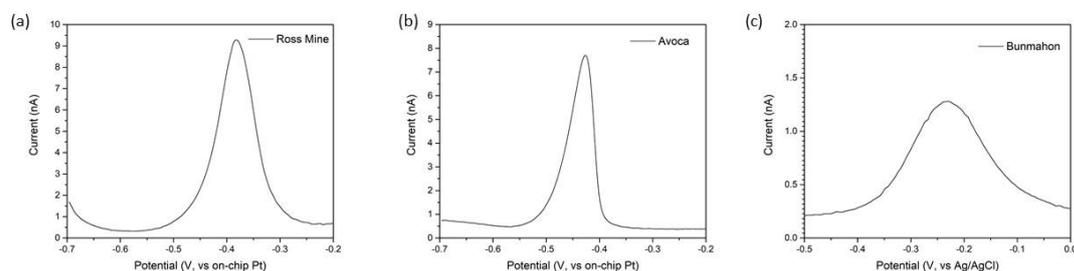


Figure 7 Square wave voltammograms of samples collected from (a) Ross mine, (b) Avoca and (c) Bunmahon (an Ag/AgCl reference electrode was used in this instance) sites. Solutions were adjusted to pH 2 on-site using mineral acid.

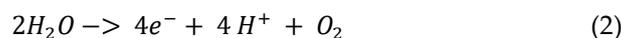
Table 2 Comparison of results obtained with ICP-MS and electrochemical sensors

| Sites | ICP-MS ($\mu\text{g/L}$) | Pt IDA Sensor ($\mu\text{g/L}$) |
|------------|----------------------------|-----------------------------------|
| Avoca | 22 | 17 |
| Ross Mines | 27 | 20 |
| Bunmahon | <3 | 1 |

3.5 Detection of Cu in standard solutions using electrochemical pH adjustment

- Electrochemical pH control

Figure 5 shows that the detection of Cu is more sensitive when the solution is at pH 2. However, typical river samples have a pH around 7. As a result, direct acidification of the sample using mineral acids is generally used for copper sensing. In the present work, the possibility to locally change the pH using electrochemical methods was investigated in order to eliminate the requirement of sample acidification prior to the analysis. The method takes advantage of the efficient catalytic properties of platinum towards the electrochemical water splitting reaction described by equation (2):



To change the pH at the surface of the sensing electrode, the protonator electrode was biased at potentials inducing reaction (2). The H^+ ions generated during this chemical reaction diffuse in the volume above the overall microelectrode (sensing electrode and protonator), resulting in a local decrease in the pH at the sensors, enabling *in situ* pH control.

To confirm the electrochemical pH control capabilities of the sensor chips, Au was electrochemically deposited on the sensing electrode. Cyclic voltammograms of 100 $\mu\text{g/L}$ Cu solutions at pH 2 to 5 in the potential range from 0.3 to 1.2 V of this modified electrode was undertaken. The peaks observed in Figure 8 (a) between 0.55V and 0.80V correspond to the reduction of the AuO formed on the forward sweep [41] and occur at 0.76 V at pH 2. pH of the solution can therefore be measured using the potential of the AuO/Au reduction peak as a metric. To test for electrochemical pH control, cyclic voltammograms of the Cu solutions at pH 7 was acquired while biasing the protonator at potential inducing the water splitting reaction. Figure 8 (b) shows the AuO reduction peak as a function of the potential applied to the protonator. The AuO/Au reduction peaks can be seen to occur at 0.76 V – indicative of a local pH of 2 – when the potential applied to the protonator was set at 1.7 V.

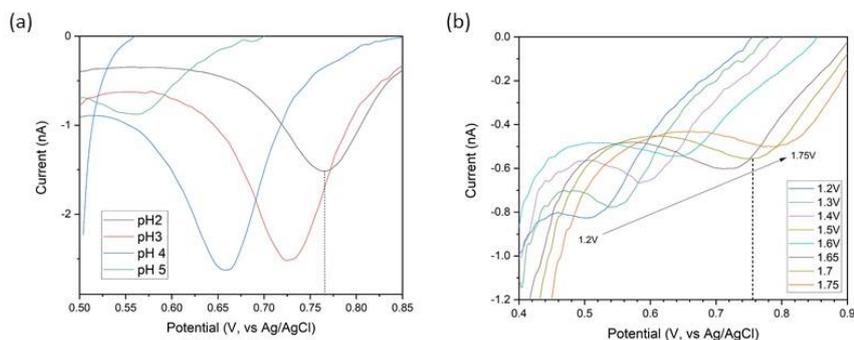


Figure 8 (a) Reverse sweep of cyclic voltammograms recorded in a 100 $\mu\text{g/L}$ Cu solution acidified to pH 2 to 5. The AuO/Au reduction peak occurs at 0.76 V at pH 2 (b) Selected regions of the cycling voltammogram of a 100 $\mu\text{g/L}$ Cu at pH 7 showing the gold oxide reduction peak when applying potentials 1.2 to 1.75V to the protonator. The dashed line shows the location of the gold oxide reduction peak maximum when 1.7 V was applied to the protonator occurring at the same position as when the solution was acidified to pH 2 with acid.

- Detection of Cu in standard solutions using electrochemical pH adjustment

To demonstrate reagent free detection of copper, the electrochemical pH control was tested on samples spiked with different concentrations of Cu. The deposition and stripping steps are depicted in Figure 9.

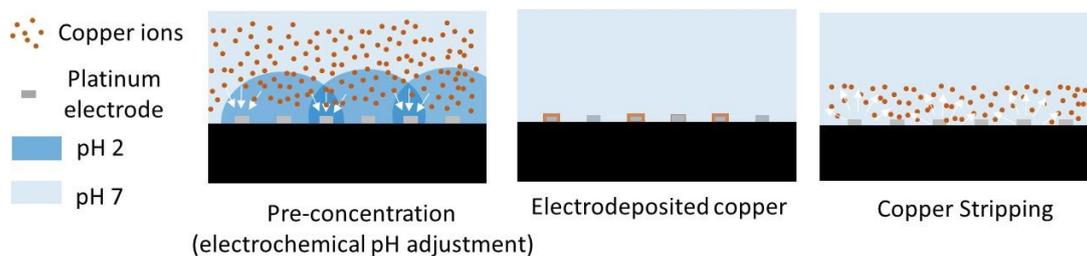


Figure 9 Schematic of detection of copper using electrochemical pH adjustment Step 1: pre-concentration at the surface of the working electrode when applying 1.7V at the protonator; step 2: stripping of deposited copper back in solution.

The protonator was biased at 1.7 V in order to obtain pH 2 at the surface of the sensing comb. The sensing comb was set at -0.4 V as previously in order to reduce the copper ions at the surface of the electrode. The square wave voltammograms of copper solutions at various concentrations can be found in Figure 10 (a). As in the case of chemical pH adjustment, the intensity of the stripping peak varies linearly with the concentration of the target analyte. Figure 10 (b) shows the corresponding calibration line between 5 and 90 $\mu\text{g/L}$. A slope of 0.29 nA/ $(\mu\text{g/L})$ was obtained in this case, suggesting the electrochemical pH control method was slightly less sensitive than the chemical pH control one. This might arise as a result of interference from dissolved oxygen also generated in the water splitting reaction.

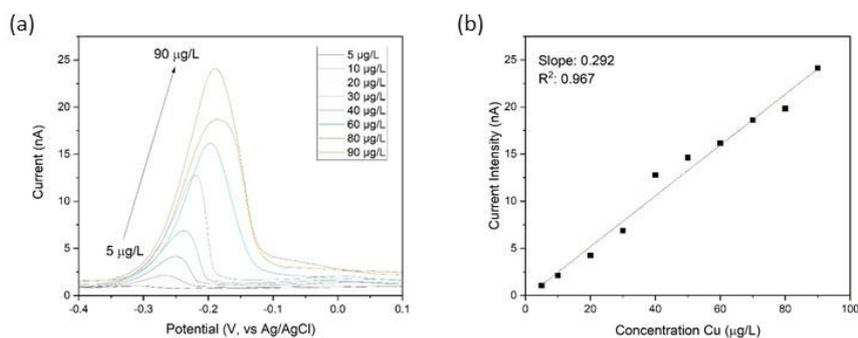


Figure 10: (a) Square wave stripping voltammograms of copper solutions acquired using electrochemical pH control. The measurements were done in copper solutions at pH 7 with 1.7 V applied at protonator IDA during deposition (b) Corresponding calibration line

Conclusions and Outlook

In summary, we have successfully fabricated platinum interdigitated microband array (IDA) on silicon substrates for trace detection of copper in water. The sensor chips comprised 8 individual sensors, allowing for multiplexed sensing, and their fabrication is compatible with mass manufacturing processes. Using square wave anodic stripping voltammetry and addition of mineral acids, the sensors exhibited high sensitivity towards detection of copper ions, with a limit of detection of 0.8 µg/L achieved. These sensors were also used for on-site testing and concentrations of 17, 20 and 1 µg/L were measured at Avoca, Ross and Bunmahon sites. ICP-MS analysis of these samples gave values of 22, 27 and < 3 µg/L, respectively and the discrepancies with the present sensors was attributed to nature of copper being detected (copper ions for electrochemical sensors vs total copper for ICP-MS). Detection of copper was also shown to be possible without addition of acids by locally changing the pH using electrochemical processes at the second comb of the IDA. Future work will now focus on developing and optimizing the electrochemical pH control method for analysis in the field.

Author Contributions: R.D: investigation, writing – original draft preparation, formal analysis; T.N: Investigation, formal analysis; A.O’R: Co supervision, funding acquisition, project administration; P.L: conceptualization, investigation, supervision, funding acquisition, writing – review and editing.

All authors have read and agreed to the published version of the manuscript.

Funding: The authors would like to acknowledge funding from the Irish EPA UisceSense Project (Code: 2015-W-MS-21). This work has been supported in part by a research grant for the VistaMilk Center Science Foundation Ireland (SFI) and Department of Agriculture Food and the Marine (DAFM) under Grant Number 16/RC//3835 as well as European Union’s Horizon 2020 research and innovation programme under the Marie Skłodowska-Curie grant agreement No: H2020- MSCA-ITN-2018-813680

Acknowledgments: We would like to thank the fabrication facility at Tyndall National Institute for their role in the production of the devices used in this research.

Conflicts of Interest: The authors declare no conflict of interest.

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