

# Selective uptake of Ag(I) from aqueous solutions using ionic liquid modified iron oxide nanoparticles

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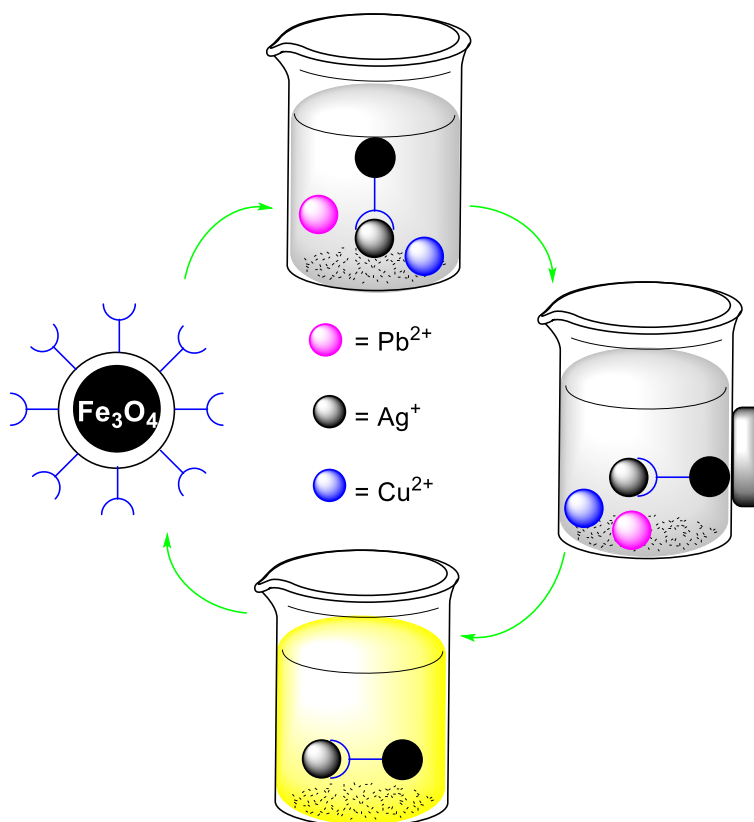
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## Graphical Abstract



## Abstract

Surface functionalized magnetic nanoparticles represent a potentially highly valuable tool for the selective recovery of metals from the aqueous phase, due to their ability to be manipulated and then recovered using an externally applied magnetic field. Ionic liquids are ideal candidates for such surface functionalization for a range of reasons, including

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16 their enhanced selectivity, low water consumption and high chemical stability. Herein the removal of  $\text{Ag}^+$  onto  
17  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  as a function of pH, exposure time, nanosorbent concentration and type of stripping  
18 agent has been investigated.  $\text{Ag}^+$  removal was recorded to fit the Langmuir isotherm indicating monolayer formation,  
19 with a saturation capacity of 23.69 mg/g. Moreover, optimum conditions for the selective removal of  $\text{Ag}^+$  in preference  
20 to  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$ , were recorded at pH 3, exposure time ranging between 0-15 min and with the highest nanosorbent  
21 concentration tested (80 mg/10ml of adsorbate solution). In addition, the most efficient stripping agent for the sorbed  
22  $\text{Ag}^+$  was determined to be thiourea at 0.6 M. Overall the results indicate that  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  is a  
23 highly adaptable and efficient agent for the selective recovery of Ag from the aqueous phase.

24 Keywords: Selective silver recovery; ionic liquid; iron oxide nanoparticle; surface functionalization; soft donor.

## 25 Introduction

26 As the global population continues to expand, demand for modern products and services which use silver (Ag),  
27 including: electronic equipment, catalysis, antibacterial agents, jewelry, water filtration media, etc., will almost  
28 certainly continue to increase (Sahan et al. 2019; Taillades and Sarradin 2004; Butterman and Hilliard 2005).  
29 Moreover Ag is listed within the EU 27 critical raw materials, and currently exhibits an “end-of-life recycling input  
30 rate” of only ~14% (Butterman and Hilliard 2005). Therefore, in order to overcome this urgent and burgeoning  
31 problem new technology is required for the enhanced recovery of Ag from our waste materials and end-of-life products  
32 (Avarmaa et al. 2019). A key challenge associated with this, however, is that such waste is typically chemically  
33 complex and mixed with a wide range of ancillary metal(oids)/materials. It is therefore clear that the development of  
34 increasingly efficient, selective and cost-effective  $\text{Ag}^+$  recovery process is highly beneficial.

35 To date, much research and development has been conducted on the removal of Ag ions from the aqueous phase  
36 including methods such as: solvent extraction (Daubinet and Kaye 2002) ion exchange (Virolainen et al. 2015)  
37 chemical precipitation (Ahlatacı et al. 2016) and solid phase extraction (Abdolmohammad-Zadeh and Javan 2015;  
38 Karimi et al. 2012). Solid phase removal media have often been preferred due to the simplicity of their application,  
39 their often low disposal costs, and often high removal efficacy (Abdolmohammad-Zadeh and Javan 2015). Within  
40 such applications, nanoparticles, defined as particles which exhibit at least one length  $<100\text{nm}$ , have received great  
41 interest due to their superior surface area and commensurate high reactivity with the aqueous phase. Such materials  
42 can also be utilized in various new applications due to their ability to be suspended in the aqueous phase as a colloid  
43 (e.g. subsurface injection). Magnetic nanoparticles have received particularly high interest due to their additional  
44 ability to be manipulated and then recovered from the aqueous phase using an externally applied magnetic field.  
45 Additional important properties for nanosorbents include stability across a wide pH range, high and rapid ion  
46 extraction efficacy, facile ion stripping efficacy and low synthesis cost. Nanosorbents possessing these desired  
47 properties have been reported. For example, in 2018, some researchers reported that the nanosorbent -  
48  $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2@ \text{Ag}^+$ -imprinted 2,4-diamine-6-phenyl-1,3,5-triazine demonstrated a higher distribution ratio  
49 and selectivity coefficient than the non-imprinted analogue in the selective extraction and preconcentration of  $\text{Ag}^+$   
50 (Jalilian and Taheri 2018). The nanosorbent could also be easily separated by a magnet and recycled. Extraction  
51 efficiency, however, was reported as relatively low at acidic pH. The synthesis of a  $\text{Ag}^+$  imprinted 3-

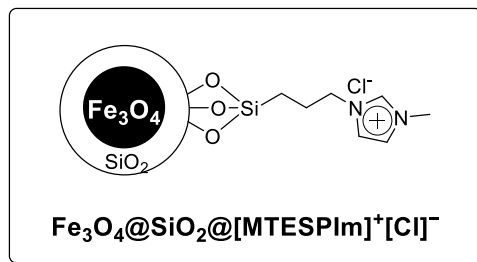
52 (triethoxysilyl)propane-1-thiol tethered to  $\text{Fe}_3\text{O}_4@\text{SiO}_2@\text{TiO}_2$  was reported and it was determined that the  
53 nanosorbent exhibits relatively high selectivity, at room temperature, for  $\text{Ag}^+$  from aqueous solutions also containing  
54  $\text{Li}^+$ ,  $\text{Co}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Ni}^{2+}$  (Yin et al. 2017). The industrial scale synthesis of such nanosorbents, however, is likely to be  
55 expensive due to the requirement for ultrapure reagents.

56 To date, research output on the synthesis and application of nanosorbents for the industrial scale selective extraction  
57 and preconcentration of trace concentrations of  $\text{Ag}^+$  remains in its infancy. Such preliminary work include the  
58 combination of 5-amino-2-thiol-1,3,4-thiadiazole and sodium dodecyl sulphate tethered by their condensation onto  
59  $\text{Fe}_3\text{O}_4/\text{Al}_2\text{O}_3$ , which was reported to extract  $\text{Ag}^+$  selectively, rapidly and quantitatively even in the presence of several  
60 order of magnitude greater concentrations of  $\text{Zn}^{2+}$ ,  $\text{Bi}^{3+}$  and  $\text{Pd}^{2+}$  (Karimi et al. 2012). Other examples include  
61  $\text{Fe}_3\text{O}_4@\text{SiO}_2@(1E,1'E)$ -1,1'-(pentane-1,5-diylbis(2,1-phenylene))bis(*N*-(3-(trimethoxysilyl)propyl)methanimine)  
62 which was recorded to selectively remove  $\text{Ag}^+$  in preference to  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  (Banaei et al. 2015). Furthermore, 1-  
63 methyl-3-[(3-trimethoxysilyl)propyl] imidazolium chloride ( $[\text{MTMSPIm}]^+[\text{Cl}]^-$ ) anchored onto  $\text{Mn}_3\text{O}_4@\text{SiO}_2$   
64 nanoparticles and demonstrated high selectivity, reusability and efficiency for extraction and preconcentration of  
65 ultratrace concentrations of  $\text{Ag}^+$  (i.e. 60 ng/mL). A key shortcoming, however, was the fact that relatively high  
66 nanosorbent mass to sample volume ratios were required. The nanosorbent was also not ideally suited for magnetic  
67 recovery applications due to the significantly lower magnetization saturation of  $\text{Mn}_3\text{O}_4$  compared to  $\text{Fe}_3\text{O}_4$  (Ozkaya et  
68 al. 2008). Ionic liquids (ILs) are compounds composed only of ions and with outstanding properties including little or  
69 very low volatility, low melting point, thermal stability and tunable hydrophilicity/hydrophobicity (Zhou et al. 2012;  
70 Seddon, Stark, and Torres 2000).

71 Herein we have built on this work by combining the proven selectivity of  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  for  $\text{Ag}^+$  recovery with  
72 the superior magnetic responsiveness of  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  (Fig. 1) (Ozkaya et al. 2008). Whilst  
73  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  has been synthesized and applied for various different applications (e.g. catalysis,  
74 medicine, printing (Qian et al. 2017; Sajjadifar, Zolfigol, and Tami 2019; Zhou et al. 2012; Yang et al. 2011; Wei et  
75 al. 2013; Azgomi and Mokhtary 2015; Garkoti, Shabir, and Mozumdar 2017)) to the best of our knowledge this is the  
76 first investigation into various parameters for the selective uptake of  $\text{Ag}^+$ . The aim of this work was therefore to  
77 investigate its behavior towards  $\text{Ag}^+$  across a range of differential constraints (namely: pH, contact time, nanosorbent  
78 dose,  $\text{Ag}^+$  recovery efficacy by different stripping agents) in order to understand  $\text{Ag}^+$  removal kinetics and mechanisms  
79 and thereby determine optimal application conditions.

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**Fig. 1** Core-shell iron oxide/imidazolium-based  $\text{Ag}^+$  nanosorbent composite nanosorbent

84



## 85 Experimental

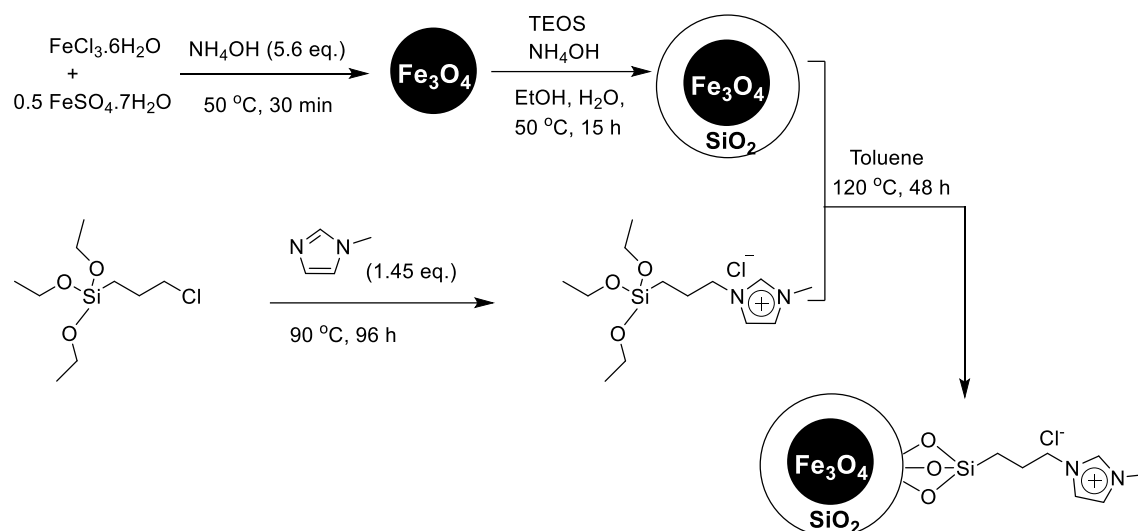
86 Chemicals.

87  $\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$  (99%),  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$  (97%),  $\text{NH}_4\text{OH}$  (25% v/v), Tetraethyl orthosilicate (99%), *N*-methylimidazole  
 88 (99%), 3-Chloropropyltriethoxysilane (97%), thiourea (99%) and toluene (99.5%) were purchased from Sigma  
 89 Aldrich, HCl (37%),  $\text{HNO}_3$  (70%),  $\text{AgNO}_3$  (99%),  $\text{NaNO}_3$  (99%) and NaOH pellets (97%) were purchased from  
 90 Fischer Scientific.  $\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$  (99%) and  $\text{Pb}(\text{NO}_3)_2$  (99%) were purchased from Acros Chemicals. Ethanol  
 91 absolute (99.8%) was purchased from VWR. All chemicals were used as received without further purification.

92 Equipment and characterization.

93 All  $^1\text{H}$  and  $^{13}\text{C}$  NMR spectra were recorded at room temperature on Bruker© Advance spectrometers. Fourier  
 94 transform infra-red (FTIR) spectra were recorded on Bruker© Alpha Platinum-Attenuated Total Reflectance IR  
 95 spectrometer. X-ray diffraction (XRD) data were collected on a Panalytical Empyrean X-ray diffractometer employing  
 96 a  $\text{Co K}\alpha$  radiation at 40 kV and 40 mA. Transmission electron micrographs (TEM) were captured by means of the  
 97 JEOL 2100+ machine operating an acceleration voltage of 200 kV from samples prepared on a copper EM grid. X-  
 98 ray photoelectron spectroscopy (XPS) data were collected on the Kratos© AXIS Ultra DLD spectrometer and take off  
 99 angle of  $90^\circ$  was used with Al(mono) x-ray source. Metal concentrations were measured by means of a PerkinElmer  
 100 5300DV Inductively Coupled Plasma Optical Emission spectrophotometer (ICP-OES). Thermogravimetry analyses  
 101 (TGA) were undertaken by means of a Metler Toledo© DSC1 – STAR at a scan rate of  $10^\circ\text{C}/\text{min}$  on samples placed  
 102 inside  $70\ \mu\text{L}$  alumina pans under a nitrogen atmosphere from  $25 - 900^\circ\text{C}$ . Magnetization data were recorded on a  
 103 quantum design MPM S5S SQUID magnetometer at 300 K. Finally, the pH of the adsorbate solutions was monitored  
 104 using the Hanna HI 8424 portable pH meter.

105 Synthesis of the ionic liquid immobilized on silica coated Fe<sub>3</sub>O<sub>4</sub> nanoparticles.



106

107 **Scheme 1** Syntheses of the nanoparticles; Fe<sub>3</sub>O<sub>4</sub> and Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>, the ionic liquid; 1-methyl-3-[(3-  
 108 triethoxysilyl)propyl] imidazolium chloride [MTESPIm]<sup>+</sup> [Cl]<sup>-</sup> and the ionic liquid-modified silica coated Fe<sub>3</sub>O<sub>4</sub>  
 109 nanosorbent.

110 The synthesis of the ionic liquid modified silica coated Fe<sub>3</sub>O<sub>4</sub> nanosorbent was achieved over four steps (Scheme 1).  
 111 In the first step, the magnetic core - Fe<sub>3</sub>O<sub>4</sub> was prepared following the method previously reported (Naka et al. 2008).  
 112 Briefly, FeCl<sub>3</sub>·6H<sub>2</sub>O (5.41 g, 0.02 mole) and FeSO<sub>4</sub>·7H<sub>2</sub>O (2.78 g, 0.01 mole) were dissolved by stirring in water (300  
 113 mL) at 50 °C. Then, NH<sub>4</sub>OH (8.52 mL, 13.20 M, 0.11 mol) was added to the iron salts solution and vigorously stirred  
 114 for 30 minutes, after which the black solids obtained were separated using a magnet. The solids were washed with  
 115 water (100 mL x 3) and EtOH (50 mL x 3). Finally, the solids were dried *in vacuo* at 70 °C to give the black solid  
 116 Fe<sub>3</sub>O<sub>4</sub>. In the second step, the silica coated Fe<sub>3</sub>O<sub>4</sub> nanoparticle was prepared following the method previously reported  
 117 (Fan et al. 2016). Briefly, to a stirred suspension of Fe<sub>3</sub>O<sub>4</sub> (0.4 g) and TEOS (0.36 g) in dry EtOH (3 mL) at 50 °C  
 118 was added a mixture of NH<sub>4</sub>OH (0.66 mL, 13.20 M), EtOH (1.2 mL) and water (0.58 mL). The reaction was left to  
 119 stir for 8 h at 50 °C and at the end of which the product suspension was left to cool to room temperature. The solids  
 120 obtained were separated with a magnet, washed with water (25 mL x 3) and finally dried *in vacuo* at 70 °C to give  
 121 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> as a black solid. In the next step, the ionic liquid; 1-methyl-3-[(3-triethoxysilyl)propyl] imidazolium  
 122 chloride ([MTESPIm]<sup>+</sup>[Cl]<sup>-</sup>) was accessed following a protocol previously reported (Abdolmohammad-Zadeh and  
 123 Javan 2015). Hence, a mixture of *N*-methylimidazole (4.8 mL, 0.06 mol) and 3-chloropropyltriethoxysilane (9 mL,  
 124 0.04 mol) was refluxed at 90°C for 96 h, after which the crude product was left to cool to room temperature. This  
 125 crude product was washed with dry diethyl ether (200 mL x 3) and dried *in-vacuo* at room temperature to give the  
 126 ionic liquid; [MTESPIm]<sup>+</sup> [Cl]<sup>-</sup> as a light brown oil. Yield: 10.27 g (53 %), <sup>1</sup>H NMR (300 MHz, DMSO-d<sub>6</sub>) δ 9.35 (s,  
 127 1H, NCHN), 7.80 (d, *J* = 6.0 Hz, 2H, NCHCHN), 4.16 (t, *J* = 6.5 Hz, 2H, CH<sub>2</sub>CH<sub>2</sub>N), 3.87 (s, 3H, CHNCH<sub>3</sub>), 3.74  
 128 (q, *J* = 6.5 Hz, 6H, CH<sub>3</sub>CH<sub>2</sub>O), 1.81 (m, 2H, CH<sub>2</sub>CH<sub>2</sub>CH<sub>2</sub>), 1.14 (t, *J* = 6.5 Hz, 9H, CH<sub>3</sub>CH<sub>2</sub>O), 0.51 (t, *J* = 6.5 Hz,  
 129 2H, SiCH<sub>2</sub>CH<sub>2</sub>CH<sub>2</sub>). <sup>13</sup>C NMR (75 MHz, DMSO-d<sub>6</sub>) δ 136.7 (NCHN) 123.6 (NCHCHN), 122.2 (NCHCHN), 57.8

130 (CH<sub>2</sub>CH<sub>2</sub>O), 51.0 (CH<sub>3</sub>CH<sub>2</sub>N), 35.7 (CHNCH<sub>3</sub>), 23.7 (CH<sub>2</sub>CH<sub>2</sub>CH<sub>2</sub>), 18.2 (CH<sub>3</sub>CH<sub>2</sub>O), 6.7 (SiCH<sub>3</sub>CH<sub>3</sub>), *m/z* (ESI)  
131 [M+Na]<sup>+</sup> 348. Finally, a previously reported method was employed to prepare the ionic liquid ([MTESPIIm]<sup>+</sup> [Cl]<sup>-</sup>)  
132 immobilized on the magnetic nanoparticle (Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>) (Chen et al. 2014). Briefly, the Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> (1.5 g) was  
133 dissolved in toluene (300 mL) and sonicated in an ultrasound bath at room temperature for 10 min. Meanwhile, the  
134 ionic liquid (15.40 g) was dissolved in toluene (100 mL) and sonicated at room temperature for 10 min. The ionic  
135 liquid solution was then added to the Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> and the mixture was refluxed at 120° C for 48 h. After 48 h, reaction  
136 mixture was left to cool to room temperature and supernatant was decanted. The solid left behind was then washed  
137 with deionized water (250 mL x 2) and EtOH (250 mL x 3) and finally dried *in vacuo* at 70°C to give the nanosorbent;  
138 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>@[MTESPIIm]<sup>+</sup>[Cl]<sup>-</sup> as a brown solid.

139 Determination of particle diameter.

140 Particle diameter was evaluated using the Scherrer equation as defined below;

$$141 \quad \tau = \frac{K\lambda}{\beta \cos\theta}$$

142 Where  $\tau$  = particle diameter,  $K$  = Scherrer constant for spherical particles (0.94),  $\lambda$  = Wavelength of X-ray source,  $\theta$  =  
143 Bragg (diffraction) angle of the most intense peak and  $\beta$  = broadening at half the maximum intensity (FWHM) of the  
144 most intense peak.

145 Determination of magnetic nanoparticle surface coverage.

146 The surface coverage of magnetic nanoparticle was evaluated using the equation defined below;

$$147 \quad \text{No of molecules per nm}^2 = \frac{W \times d_{\text{Fe}_3\text{O}_4} \times r \times N_A}{M(1-W) \times 3 \times 10^{21}}$$

148 Where  $W$  is the weight loss of sample,  $d_{\text{Fe}_3\text{O}_4}$  is the density of the Fe<sub>3</sub>O<sub>4</sub> = 5.17 g/cm<sup>3</sup>,  $N_A$  is the Avogadro's constant  
149 = 6.022 x 10<sup>23</sup>,  $M$  is the molecular weight of the ligand,  $r$  is the radius of the composite nanoparticle.

150 Determination of sorption isotherms.

151 The sorption isotherm governing the removal of Ag<sup>+</sup> was predicted by means of three isotherms models - Langmuir,  
152 Freundlich and Temkin using data generated from contacting the same amount (10 mg) of the nanosorbent with  
153 varying Ag<sup>+</sup> concentrations (4 to 90 mgL<sup>-1</sup>) at pH 3, for 15 mins at room temperature. Equations representing the  
154 sorption isotherms are;

155 Langmuir isotherm:

$$156 \quad \frac{C_e}{Q_e} = \frac{1}{Q_m K_L} + \frac{C_e}{Q_m}$$

157 Where  $C_e$ ,  $Q_e$ ,  $Q_m$  and  $K_L$  are the equilibrium Ag<sup>+</sup> concentration (mgL<sup>-1</sup>), the amount of Ag<sup>+</sup> on the nanosorbent (mgg<sup>-1</sup>)  
158 <sup>1</sup>), the maximum capacity of the nanosorbent (mgg<sup>-1</sup>) and the Langmuir adsorption constant (L/mg) respectively.

159 Freundlich isotherm:

$$160 \log Q_e = \log K_F + \frac{1}{n} \log C_e$$

161 Where  $Q_e$  and  $C_e$  have been described above,  $K_F$  and  $n$  are Freundlich constants related to maximum sorption capacity  
162 (mg/g) and heterogeneity factor ( $\text{mg}^{-1}$ ).

163 Temkin isotherm:

$$164 Q_e = B \ln A_T + B \ln C_e$$

165  $A_T$  ( $\text{Lg}^{-1}$ ) and  $B$  ( $\text{Jmol}^{-1}$ ) are Temkin constants related to the binding constant and heat of sorption respectively.

166  $\text{Ag}^+$  removal studies.

167 For the control study, a 500 mL aqueous solution containing  $1 \text{ mgL}^{-1} \text{ Ag}^+$  in  $0.023 \text{ M NaNO}_3$  was prepared from a  
168  $500 \text{ mgL}^{-1}$  stock solution. Afterwards, 20 mg of nanosorbents ( $\text{Fe}_3\text{O}_4$ ,  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  and  
169  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) were separately contacted with 10 mL of the  $\text{Ag}^+$  aqueous metal solution at pH 1  
170 inside 30 mL plastic screw cap vials. After 45 min the solids were magnetically separated in about 2 min and the  
171 supernatant was removed using a plastic syringe and prepared for metal content determination by ICP-OES. For the  
172 adsorption isotherm study, 10 mg of the nanosorbent ( $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) was exposed to 10 mL  
173 aqueous solutions containing varying  $\text{Ag}^+$  concentrations ( $4$  to  $90 \text{ mgL}^{-1}$ ) in  $0.023 \text{ M NaNO}_3$  at pH 3 for 15 min. The  
174 batch experiments used to investigate the optimum conditions for the selective removal of  $\text{Ag}^+$  onto  
175  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  were conducted by exposing 10 mg of the nanosorbent to a 10 mL aqueous solution  
176 containing  $\text{Cu}^{2+}$ ,  $\text{Ag}^+$  and  $\text{Pb}^{2+}$  each at  $2 \text{ mgL}^{-1}$  in  $0.023 \text{ M NaNO}_3$  at an initial pH of 3 unless otherwise stated. The  
177 pH of the metal aqueous solutions was adjusted to the desired pH using a few drops (typically between 1 and 10) of  
178 dilute  $0.001 \text{ M HNO}_3$  or  $\text{NaOH}$ . For initial pH study, pH of adsorbate solution was varied from 1 to 5. For the contact  
179 time study, contact time was varied from 0 to 90 mins. For nanosorbent dose study, amount of nanosorbent used was  
180 varied from 5 to 80 mg. All batch studies were undertaken in triplicates and at room temperature ( $21 \text{ }^\circ\text{C}$ ).

181 The metal removal efficiency was determined by the equation below;

$$182 \% RE = \frac{C_i - C_f}{C} \times 100$$

183 Where  $\% RE$  is percentage removal efficiency,  $C_i$  and  $C_f$  (in  $\text{mgL}^{-1}$ ) are the initial and final metal ion concentrations  
184 respectively.

185 The selectivity factor was determined by the following equation (Shamsipur et al. 2014);

$$186 K_{Ag^+/M^{n+}} = \frac{K_d^{Ag^+}}{K_d^{M^{n+}}}$$

187 
$$K_d = \frac{(C_i - C_f)v}{mC_f}$$

188  $K_d$  is distribution ratio,  $C_i$  and  $C_f$  are initial and final metal ion concentration respectively (in  $\text{mgL}^{-1}$ ),  $v$  is the volume  
189 of aqueous solution in mL,  $m$  is the mass of nanosorbent (in mg).

190  $\text{Ag}^+$  stripping efficiency studies.

191 For the stripping efficiency study,  $\text{Ag}^+$ -impregnated  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  nanosorbents were washed  
192 with de-ionized water and magnetically separated after which it was contacted with 5 mL of stripping agent (HCl,  
193  $\text{HNO}_3$  or thiourea) inside screw-capped plastic vials for 1 h. Thereafter, the nanosorbent were magnetically separated  
194 and the supernatant stripping agent solution was carefully withdrawn using a plastic syringe. The stripping agent  
195 solution was made up to 10 mL by adding deionized water and the metal content was determined again by ICP-OES.  
196 The experiments were undertaken in triplicates. The stripping efficiency of  $\text{Ag}^+$  by a stripping agent was determined  
197 following the equation;

198 
$$\%SE = \frac{C_e}{C_i} \times 100$$

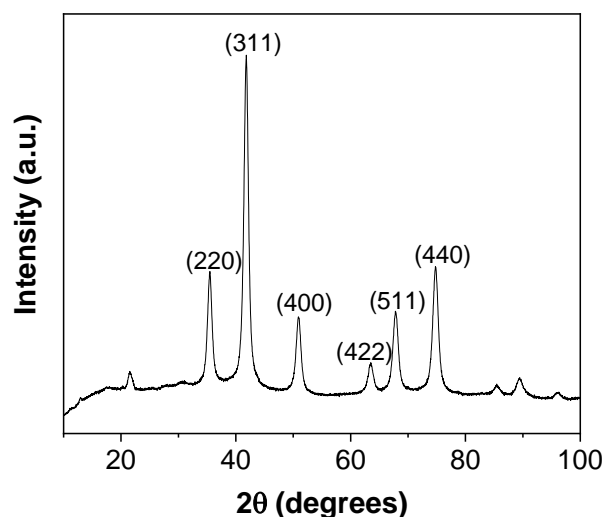
199 Where  $\%SE$  is percentage stripping efficiency,  $C_e$  represents the concentration of  $\text{Ag}^+$  recovered by the stripping agent  
200 ( $\text{Ag}^+$  concentration in the stripping agent solution) and  $C_i$  represents the initial  $\text{Ag}^+$  concentration respectively  
201 (concentration of  $\text{Ag}^+$  in nanosorbent prior to stripping).

## 202 **Results and Discussion**

203 Physical and chemical characterization of  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ .

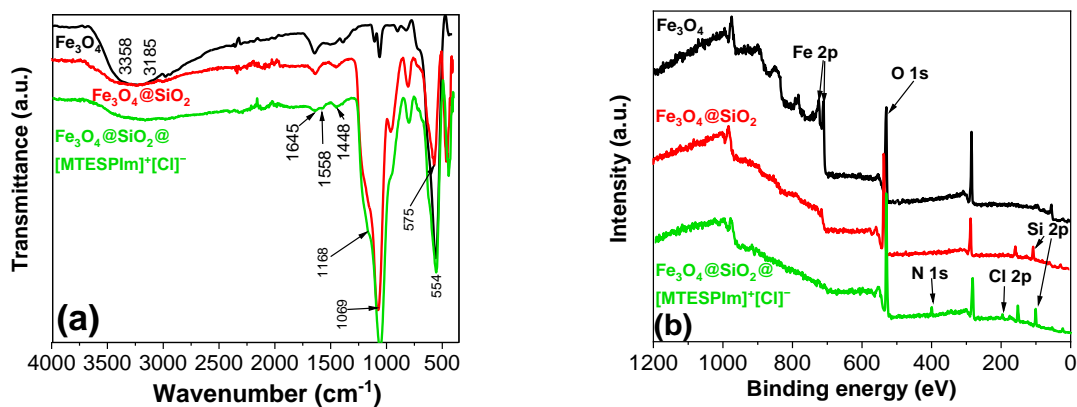
204 The x-ray diffractogram of the black powder obtained after the treatment of  $\text{FeSO}_4$  and  $\text{FeCl}_3$  with aqueous ammonia  
205 is presented in Fig. 2. The experimental  $d$ -spacings from the x-ray diffractogram of the  $\text{Fe}_3\text{O}_4$  are 2.96, 2.52, 2.09,  
206 1.71, 1.61, 1.48 at  $2\theta$  ( $^\circ$ ) of 35.24, 41.58, 50.69, 63.24, 67.57 and 74.52 respectively. While the  $d$ -spacings are  
207 characteristic of  $\text{Fe}_3\text{O}_4$  and identical to those reported in literature (Naka et al. 2008; Sun et al. 2007), the  $2\theta$  values  
208 are not, attributed to the difference in the x-ray sources used – Co  $K\alpha$  in this study but Cu  $K\alpha$  in the literature (Naka  
209 et al. 2008; Sun et al. 2007). The average diameter of the  $\text{Fe}_3\text{O}_4$  particles evaluated by the Scherrer equation (Puig et  
210 al. 2012) was found to be  $10.0 \pm 0.3$  nm.





211  
 212 **Fig. 2** X-ray diffractogram of  $\text{Fe}_3\text{O}_4$  nanoparticles (Co  $K\alpha$  source wavelength = 1.79 Å, voltage = 40 kV, current =  
 213 40 mA)

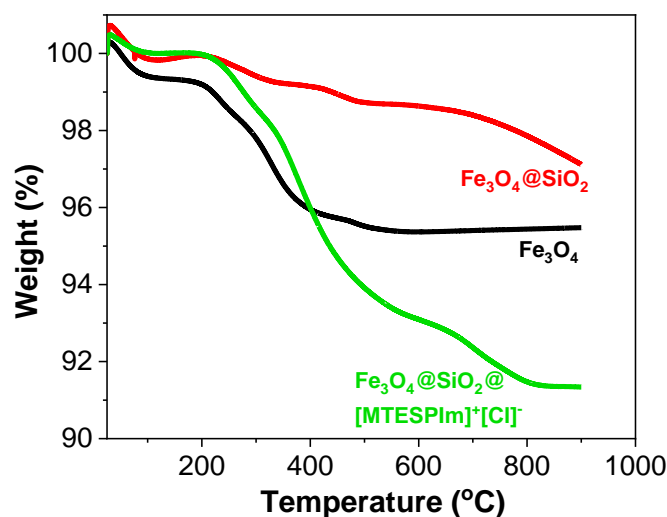
214 The chemical composition of the different nanoparticles was each characterized using FTIR and XPS. FTIR spectra  
 215 of all three nanoparticles (Fig. 3a) contain a peak around  $554\text{ cm}^{-1}$  which was attributed to the Fe-O vibrations in  
 216  $\text{Fe}_3\text{O}_4$  (Abbas et al. 2014). The spectrum for  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  also contained an intense peak at  $1069\text{ cm}^{-1}$  (attributed to  
 217 the Si-O-Si stretching vibrations) and suggesting that  $\text{SiO}_2$  has been chemisorbed onto  $\text{Fe}_3\text{O}_4$  (Abbas et al. 2014;  
 218 Sajjadifar, Zolfigol, and Tami 2019). While the spectra of the  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  and the  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$   
 219 may look identical, the low intensity peak at  $1558\text{ cm}^{-1}$  (attributed to the -C=N stretching vibration) slightly  
 220 differentiates them, indicating that the ionic liquid ( $[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) may have been linked to the  $\text{Fe}_3\text{O}_4@\text{SiO}_2$   
 221 (Minsik Kim, Hwang, and Yu 2007).



222  
 223 **Fig. 3** (a) FTIR and (b) XPS spectra of  $\text{Fe}_3\text{O}_4$  (black),  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  (red) and  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$   
 224 (green)

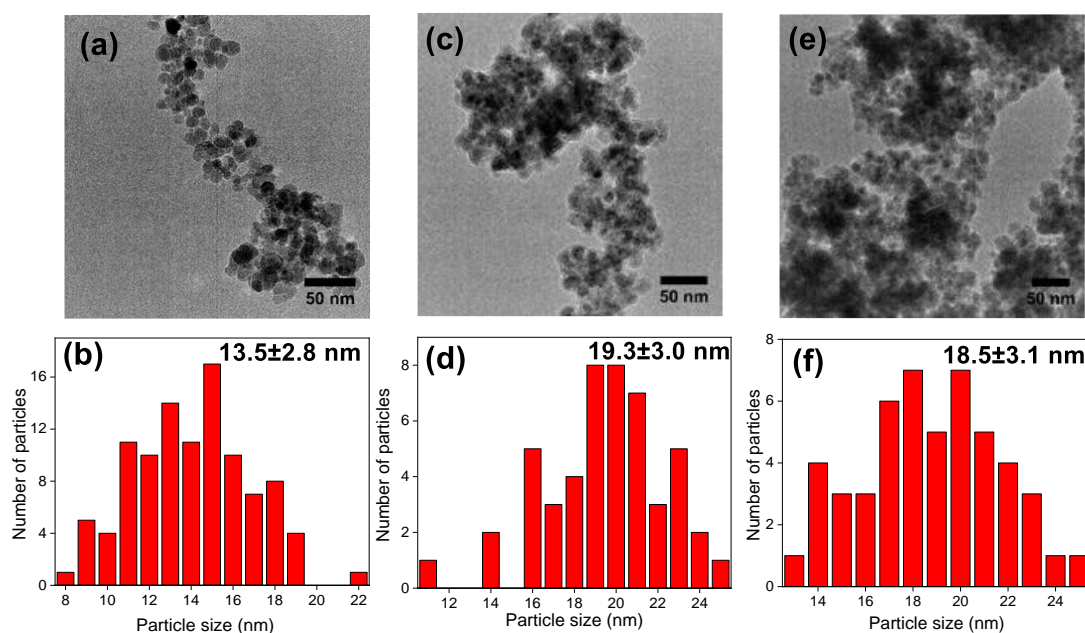
225 More information allowed us to confirm the formation of the desired  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ . The XPS  
226 spectra (Fig. 3b) confirms the FTIR data about the chemical compositions of all three nanoparticles. For example,  
227 photoelectron lines representing Fe 2p and O 1s at 725/710 and 533 eV respectively can be observed in the survey  
228 spectra of all three nanoparticles (Sun et al. 2007). Furthermore, the spectra of the  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  is clearly different  
229 from that of the  $\text{Fe}_3\text{O}_4$  with the presence of the extra photoelectron line representing Si 2p at 106 eV, confirming that  
230  $\text{SiO}_2$  has been chemisorbed onto the  $\text{Fe}_3\text{O}_4$ . Moreover, the spectra of the  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  (Fig. 3b  
231 (green)) also contains the Si 2p photoelectron line in addition to those of the N 1s, C 1s and Cl 2p at 400, 284.8 and  
232 199 eV respectively, confirming the linkage of the ionic liquid to the  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  (Sun et al. 2007; Korin, Froumin,  
233 and Cohen 2017). The photoelectron lines at 284.8 eV (characteristic of the C 1s) in the spectra of all nanoparticles;  
234  $\text{Fe}_3\text{O}_4$ ,  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  and the  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  nanoparticles is attributed to adventitious carbon  
235 (Munho Kim et al. 2017; Miller, Biesinger, and McIntyre 2002).

236 The difference in surface compositions of the nanoparticles was further highlighted by the different weight loss  
237 patterns exhibited by the different nanoparticles in the TGA thermogram (Fig. 4). The weight loss of 4 % for the  $\text{Fe}_3\text{O}_4$   
238 nanoparticle between 180 and 570 °C was attributed to the loss of trapped water molecules in the  $\text{Fe}_3\text{O}_4$  lattice and  
239 perhaps adventitious carbon as well (Khoobi et al. 2015). Expectedly, the silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles  
240 ( $\text{Fe}_3\text{O}_4@\text{SiO}_2$ ) remained stable losing only 3 % in the entire experimental temperature range. The ionic liquid-  
241 modified silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles ( $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) lost only 9 % of its weight attributed  
242 to the decomposition of the ionic liquid coating (Xu et al. 2013). The surface coverage of the ionic liquid coating on  
243 the  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  was found to be approximately 2 molecules/nm<sup>2</sup>.



244  
245 **Fig. 4** TGA thermograms for  $\text{Fe}_3\text{O}_4$  (black), silica coated  $\text{Fe}_3\text{O}_4$  (red) and ionic liquid-modified silica coated  $\text{Fe}_3\text{O}_4$   
246 nanoparticles (green) (Atmosphere: nitrogen, heating rate: 10 °C/min)

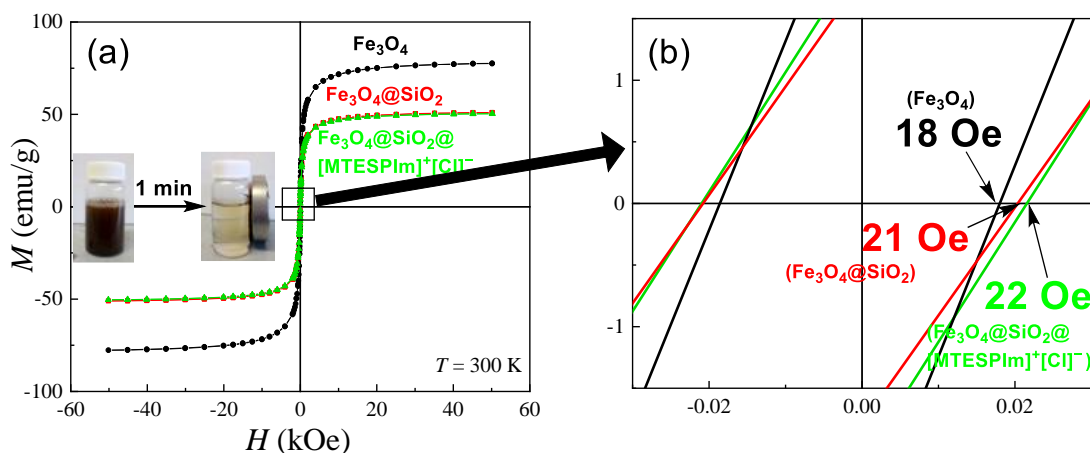
247 Under the electron microscope, the  $\text{Fe}_3\text{O}_4$  are spherical and aggregated particles with an average size of  $13.5 \pm 2.8$  nm  
 248 (Figs. 5a and b). Aggregation of bare  $\text{Fe}_3\text{O}_4$  nanoparticles is not uncommon and it has been explained that bare  $\text{Fe}_3\text{O}_4$   
 249 particles aggregate in order to reduce their surface energy (Ditsch et al. 2005). The difference between the particle  
 250 diameters obtained after analyses of the x-ray diffractogram ( $10.0 \pm 0.3$  nm) and micrograph ( $13.5 \pm 2.8$  nm) of  $\text{Fe}_3\text{O}_4$   
 251 was attributed to error associated with manual sizing of the particles from the TEM micrograph. As expected, the  
 252  $\text{Fe}_3\text{O}_4@ \text{SiO}_2$  (Fig. 5c) and  $\text{Fe}_3\text{O}_4@ \text{SiO}_2@ [\text{MTESPI}m]^+[\text{Cl}]^-$  particles (Fig. 5e) are bigger than the  $\text{Fe}_3\text{O}_4$  with average  
 253 sizes of  $19.3 \pm 3.0$  nm (Fig. 5d) and  $18.5 \pm 3.1$  nm (Fig. 5f) respectively, noting that these diameters are identical considering  
 254 their errors. The aggregation observed for the  $\text{Fe}_3\text{O}_4@ \text{SiO}_2$  (Fig. 5c) particles has been attributed to the increase in  
 255 ionic strength of the reaction medium as a result of the hydrolysis and condensation of the TEOS units (Philipse, Van  
 256 Bruggen, and Pathmamanoharan 1994). On the other hand, aggregation observed for the  
 257  $\text{Fe}_3\text{O}_4@ \text{SiO}_2@ [\text{MTESPI}m]^+[\text{Cl}]^-$  particles (Fig. 5e) may have been caused by intermolecular electrostatic attraction  
 258 between surfaces bearing the ionic liquid. Based on the sizes of the  $\text{Fe}_3\text{O}_4$  and the  $\text{Fe}_3\text{O}_4@ \text{SiO}_2$ , the average thickness  
 259 of the silica layer, which appear as grey fringes in the micrograph of the  $\text{Fe}_3\text{O}_4@ \text{SiO}_2$  (Fig. 5c) will have been about  
 260 6 nm.



261  
 262 **Fig. 5** (a,c,e) Transmission electron micrographs of (a)  $\text{Fe}_3\text{O}_4$  (c) silica coated  $\text{Fe}_3\text{O}_4$  and (e) ionic liquid-modified  
 263 silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles, (b,d,f) Histograms showing size distribution of (b)  $\text{Fe}_3\text{O}_4$  (d) silica coated  $\text{Fe}_3\text{O}_4$   
 264 and (f) ionic liquid-modified silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles.

265 The observed particle diameters indicated that all three nanoparticles are superparamagnetic (being less than 20 nm)  
 266 (Wahajuddin, A. and Arora 2012; Neamtu and Verga 2011). Indeed, this inference was confirmed from the  
 267 magnetization curves (Fig. 6a) obtained after SQUID characterization, where all three nanoparticles had low  
 268 coercivities (Mahdavian and Mirrahimi 2010; Salviano et al. 2018) ( $\text{Fe}_3\text{O}_4 = 18$ ,  $\text{Fe}_3\text{O}_4@ \text{SiO}_2 = 21$  and

269  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^- = 22 \text{ Oe}$  (Fig. 6b). Expectedly, the  $\text{Fe}_3\text{O}_4$  had the highest magnetization saturation  
 270 ( $M_s$ ) value of  $77.60 \pm 0.10 \text{ emu/g}$ . The  $M_s$  of the  $\text{Fe}_3\text{O}_4@SiO_2$  and the  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  are not  
 271 significantly different with values of  $50.99 \pm 0.01$  and  $50.30 \pm 0.10 \text{ emu/g}$  respectively, showing that the magnetic  
 272 response of the  $\text{Fe}_3\text{O}_4@SiO_2$  is not significantly reduced after the surface modification. The lower  $M_s$  for  $\text{Fe}_3\text{O}_4@SiO_2$   
 273 and  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  is attributed to the surface modifications by non-magnetic materials – silica and  
 274 the ionic liquid ( $[MTESPIIm]^+[Cl]^-$ ) respectively (Chen et al. 2014). Notwithstanding, the nanoparticle  
 275  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  can be quickly separated from an aqueous solution in about 1 min (Fig. 6a inset).



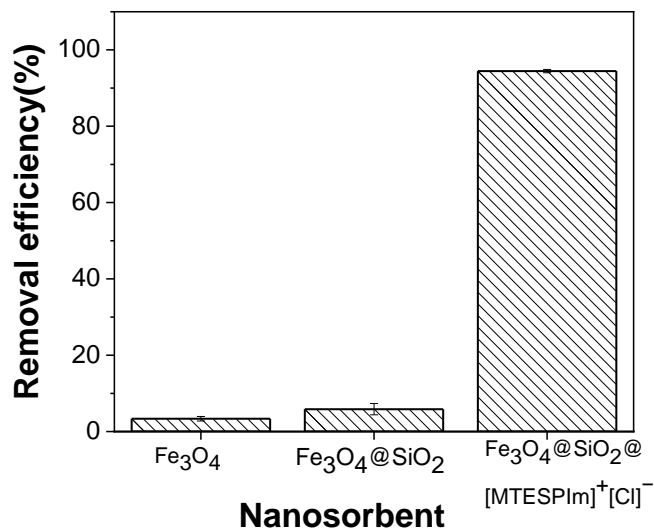
276  
 277 **Fig. 6** (a) Magnetization curves for  $\text{Fe}_3\text{O}_4$  (black), silica coated  $\text{Fe}_3\text{O}_4$  ( $\text{Fe}_3\text{O}_4@SiO_2$ ) (red) and ionic liquid-modified  
 278 silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles ( $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$ ) (green) (Inset: separation of  
 279  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  from an aqueous solution using a magnet). (b) Expanded magnetization curve  
 280 showing coercivities of  $\text{Fe}_3\text{O}_4$  (black),  $\text{Fe}_3\text{O}_4@SiO_2$  (red) and  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  (green).

281  
 282 Control study.

283 Prior to the investigation of the optimum conditions for  $\text{Ag}^+$  removal by the  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$ , it was  
 284 necessary to establish the ionic liquid;  $[MTESPIIm]^+[Cl]^-$  as the agent responsible for the  $\text{Ag}^+$  removal. Therefore, the  
 285  $\text{Ag}^+$  removal efficiencies of the nanoparticles;  $\text{Fe}_3\text{O}_4$ ,  $\text{Fe}_3\text{O}_4@SiO_2$  and  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  were  
 286 measured by contacting 20 mg of each nanoparticle with 10 mL aqueous solution containing  $1 \text{ mgL}^{-1} \text{ Ag}^+$  at pH 1 and  
 287 at room temperature.

288 The  $\text{Ag}^+$  removal by the three nanoparticles increased in the order;  $\text{Fe}_3\text{O}_4$  ( $3.4 \pm 0.6 \%$ ) <  $\text{Fe}_3\text{O}_4@SiO_2$  ( $5.8 \pm 1.5 \%$ ) <  
 289  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  ( $94.4 \pm 0.4 \%$ ), with  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  demonstrating the highest  
 290 removal efficiency (Fig. 7). The quantitative removal of  $\text{Ag}^+$  by the  $\text{Fe}_3\text{O}_4@SiO_2@[MTESPIIm]^+[Cl]^-$  may be  
 291 attributed to the preferential binding of the soft  $N$ - donor in the ionic liquid;  $[MTESPIIm]^+[Cl]^-$  to the soft  $\text{Ag}^+$  acceptor

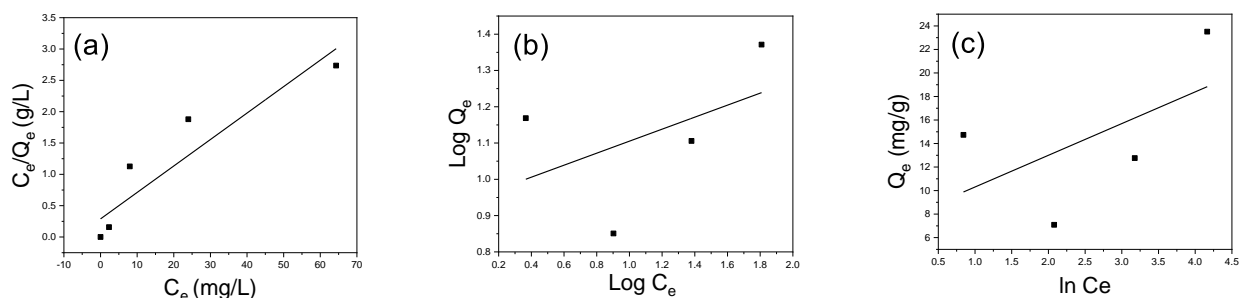
292 (Pearson 1968). Going forward, the nanoparticle  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  was employed for the investigation  
293 of the optimum conditions for the removal of  $\text{Ag}^+$ .



294  
295 **Fig. 7** Extraction efficiencies of  $\text{Fe}_3\text{O}_4$ , silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles ( $\text{Fe}_3\text{O}_4@\text{SiO}_2$ ) and ionic liquid-modified  
296 silica coated  $\text{Fe}_3\text{O}_4$  nanoparticles ( $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) for  $\text{Ag}^+$  extraction from aqueous solution.  
297 (conditions:  $[\text{Ag}^+]_0 = 1 \text{ mgL}^{-1}$ , volume = 10 mL, pH = 1.0, contact time = 45 min, temperature = RT, nanosorbent  
298 dose = 20 mg).

299  
300 Adsorption isotherms.

301 Figs. 8a-c display the sorption data plotted against Langmuir, Freundlich and Temkin isotherms after undertaking the  
302 relevant calculations. It can be observed that the Langmuir isotherm provided the best fit, with an  $R^2$  value of 0.87  
303 compared to 0.48 for Freundlich and 0.32 for Temkin (Table 1). This indicates that  $\text{Ag}^+$  removal proceeded  
304 predominantly *via* chemisorption with the formation of a monolayer on  $[\text{MTESPIIm}]^+[\text{Cl}]^-$ , with a saturation capacity  
305 of 23.69 mg/g. This agrees with Fig. 7 where the removal of  $\text{Ag}^+$  onto  $\text{Fe}_3\text{O}_4$  and  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  was comparatively  
306 much lower than  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  and therefore suggests that the majority of  $\text{Ag}^+$  sorption was with  
307 the ionic liquid. This therefore provides evidence that  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  is a potentially reusable  
308 agent for  $\text{Ag}^+$  removal.



309

310 **Fig. 8** (a) Langmuir, (b) Freundlich and (c) Temkin sorption isotherm plots for the removal of  $\text{Ag}^+$  by the ionic  
311 liquid-modified silica coated  $\text{Fe}_3\text{O}_4$  nanosorbent (pH = 3, time = 15 min, nanosorbent dose = 10 mg)

312

313 **Table 1.** Langmuir, Freundlich and Temkin isotherm parameters for the removal of  $\text{Ag}^+$  by the ionic liquid-modified  
314 silica coated  $\text{Fe}_3\text{O}_4$  nanosorbent ( $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ )

Langmuir			Freundlich			Temkin		
$Q_m$ (mg/g)	$K_L$ (L/mg)	$R^2$	$K_F$ (mg/g)	n	$R^2$	B (J/mol)	$A_T$ (L/g)	$R^2$
23.69	0.15	0.87	8.70	6.04	0.48	2.70	16.80	0.32

315

316 Effects of initial pH, contact time and nanosorbent dose on the efficiency and selectivity of removal  $\text{Ag}^+$  by the  
317 nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ .

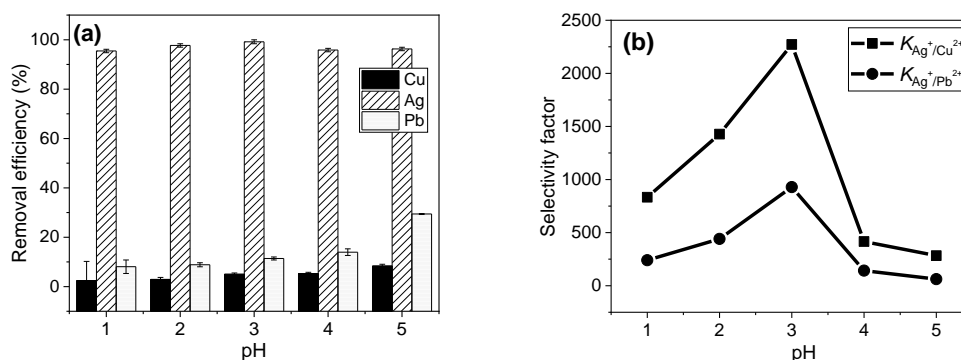
318 The effect of some conditions (including initial pH, contact time and nanosorbent dose) on the removal efficiency and  
319 selectivity of the nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  for  $\text{Ag}^+$  removal from aqueous solutions also  
320 containing competing ions ( $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$ ) were investigated by contacting the nanosorbent with 10 mL of aqueous  
321 solutions containing ca.  $2 \text{ mgL}^{-1}$  each of  $\text{Cu}^{2+}$ ,  $\text{Ag}^+$  and  $\text{Pb}^{2+}$ . The choice of  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  as competing ions was  
322 informed by the knowledge that  $\text{Ag}^+$ , usually coexists with  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  in ores and mine tailings, for example (Crane  
323 et al. 2017). Also, it was stated that very low concentrations of  $\text{Ag}^+$ ,  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  was employed in this study because  
324  $\text{Ag}^+$ , typically exists in very low concentrations in potential  $\text{Ag}^+$  repositories of interest.

325 Effect of initial pH.

326 The pH is an important parameter usually investigated in metal removal studies since metal extraction efficiencies  
327 have usually been found to be dependent upon pH (Abdolmohammad-Zadeh and Javan 2015; Karimi et al. 2012). In  
328 this study, the effect of selectivity for and extraction efficiency of  $\text{Ag}^+$  was investigated by varying the pH of the

329 mixed metal aqueous solution from 1 to 5. This pH range was chosen since the efficiency of the nanosorbent at low  
 330 pH was of particular interest because of low pH environments the nanosorbent may be applied to. Interestingly, it was  
 331 observed that the extraction efficiency of  $\text{Ag}^+$  by the nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  was dependent  
 332 on pH. The highest extraction efficiency for  $\text{Ag}^+$  (99.2±0.8 %) was obtained at pH 3 (Fig. 9a). Similar studies have  
 333 reported excellent extraction efficiencies at this pH (Daubinet and Kaye 2002; Abdolmohammad-Zadeh and Javan  
 334 2015). The quantitative removal of  $\text{Ag}^+$  at this pH (pH 3) has been explained by the hard soft acid base (HSAB) rule  
 335 (Pearson 1968). This explanation also helps to understand the quantitative recoveries of  $\text{Ag}^+$  at lower pH (pH 1 =  
 336 95.5±0.7 %, pH 2 = 97.7±0.7 %) (Fig. 9a). Furthermore, observation of quantitative extraction efficiencies of  $\text{Ag}^+$  at  
 337 very low pH indicates that the nanosorbent is likely to be stable at low pH.

338 The selectivity for  $\text{Ag}^+$  over  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  by the nanosorbent was also observed to be pH dependent. The highest  
 339 selectivity for  $\text{Ag}^+$  over both  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  was observed at pH 3 ( $K_{\text{Ag}^+/\text{Cu}^{2+}} = 2272.3$  and  $K_{\text{Ag}^+/\text{Pb}^{2+}} = 928.7$ ) (Fig. 9b).  
 340 It can be observed that the selectivity factor of the adsorbent for  $\text{Cu}^{2+}$  was higher than that for  $\text{Pb}^{2+}$  over the pH range  
 341 investigated. It is unclear why this was observed. In hard/soft terms  $\text{Cu}^{2+}$  is softer than  $\text{Pb}^{2+}$ , therefore the selectivity  
 342 for  $\text{Ag}^+$  over  $\text{Pb}^{2+}$  should be higher than over  $\text{Cu}^{2+}$  and not the other way around as observed. Other factors such as  
 343 differential removal by precipitation as Pb and Cu chlorides or electrostatic attraction of the aqua ions;  $\text{Pb}_4(\text{OH})_4^{4+}$   
 344 and  $\text{Cu}(\text{H}_2\text{O})_6^{2+}$  to the  $\text{Cl}^-$  in the nanoparticle are unlikely explanations as they both form chlorides and aqua ions  
 345 readily. It might be related to the average pore size of the nanoparticles (the ionic radius of  $\text{Pb}^{2+}$  is bigger than  $\text{Cu}^{2+}$   
 346 (1.27 vs 0.72 Å), but this couldn't be verified as the pore size of the nanoparticles were not measured. Summarily, the  
 347 highest extraction efficiency and highest selectivity of  $\text{Ag}^+$  were obtained the nanosorbent;  
 348  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  were obtained at pH 3.



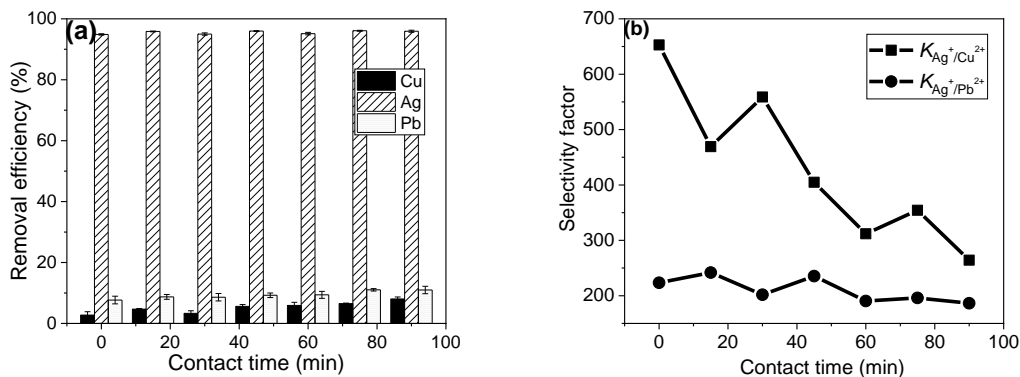
349  
 350 **Fig. 9** Effect of pH on (a) efficiency and (b) selectivity of  $\text{Ag}^+$  removal from simulated mixed metal aqueous  
 351 solution by the ionic liquid-modified silica coated  $\text{Fe}_3\text{O}_4$  nanosorbent ( $[\text{M}^{n+}] = 2 \text{ mgL}^{-1}$ , volume = 10 mL, contact  
 352 time = 2 h, temperature = RT, nanosorbent dose = 10 mg)

353 Effect of contact time.

354 The optimum contact time for the extraction of  $\text{Ag}^+$  by the nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  was  
355 investigated by varying the contact time from 0 to 90 min. A duration of 0 min refers to the situation where the  
356 nanoparticle is separated immediately after the nanosorbent had been contacted with the mixed metal aqueous solution.  
357 Usually, the magnetic separation of the nanosorbent occurs in about 1 min.

358 The extraction efficiency showed no significant difference across the contact time; 15-90 min, with the highest and  
359 lowest extraction efficiencies being  $96.1\pm 0.7\%$  (75 min) and  $94.9\pm 0.7\%$  (0 min) respectively (Fig. 10a). This  
360 observation was attributed to the availability of enough sites on the surface of the nanosorbent for the removal of  $\text{Ag}^+$   
361 with a rapid reactivity (noting even at 0 minutes 94.9% of  $\text{Ag}^+$  was removed) (Lasheen et al. 2014; Beigzadeh and  
362 Moeinpour 2016).

363 The selectivity factor of  $\text{Ag}^+$  over the interfering ions ( $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$ ) worsens over an increase in the contact time.  
364 The highest selectivity factor for  $\text{Ag}^+$  over  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  were observed at 0 and 15 min respectively (Fig. 10b), with  
365 the implication that selective  $\text{Ag}^+$  removal is favourable for the least duration possible suggesting a kinetically  
366 controlled selectivity. The lowest selectivity factor of  $\text{Ag}^+$  over  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  were observed at 90 min, indicating  
367 that exchange of metals may be occurring in line with a thermodynamic equilibrium being reached, thus allowance  
368 for a lengthy contact time reduces the selective extraction of  $\text{Ag}^+$ .



369  
370 **Fig. 10** Effect of contact time on (a) efficiency and (b) selectivity on  $\text{Ag}^+$  removal from simulated mixed metal  
371 aqueous solution using the ionic liquid-modified silica coated  $\text{Fe}_3\text{O}_4$  nanosorbent (Conditions:  $[\text{M}^{n+}] = 2 \text{ mgL}^{-1}$ ,  
372 volume = 10 mL, pH = 1.30, temperature = RT, nanosorbent dose = 10 mg).

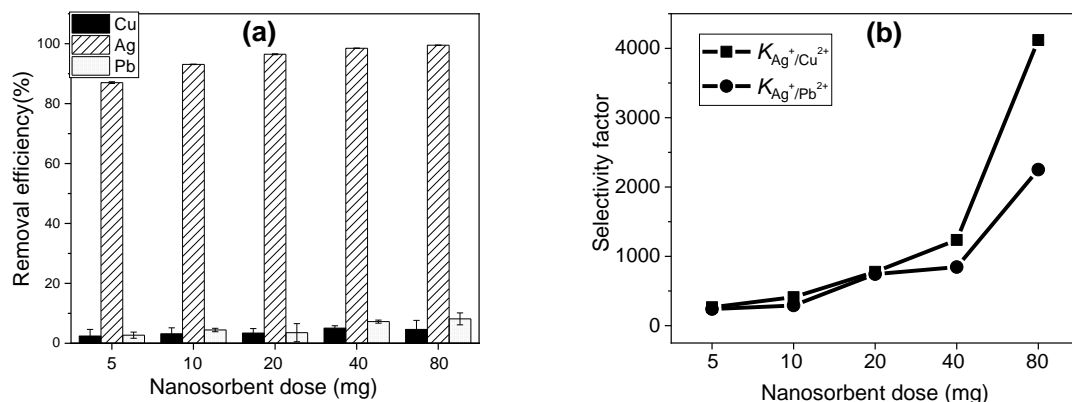
373 Effect of nanosorbent dose.

374 In order to develop an efficient and cost-effective metal extraction system, determination of the optimum dose of the  
375 nanosorbent needed is important. Therefore, the effect of nanosorbent ( $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ ) dose on  
376 efficiency and selectivity of  $\text{Ag}^+$  was investigated by varying the amount of the nanosorbent contacted with the mixed  
377 metal solution from 5 to 80 mg for 45 min. Unsurprisingly, the extraction efficiency for  $\text{Ag}^+$  extraction increased with



378 increasing nanosorbent dose (Fig. 11a). The lowest and highest extraction efficiencies for  $\text{Ag}^+$  were observed at the  
 379 lowest and highest amounts of nanosorbent used ( $87.0 \pm 0.3\%$  at 5 mg and  $99.5 \pm 0.1\%$  for 80 mg). The increase of  
 380  $\text{Ag}^+$  extraction efficiency with increasing nanosorbent dosage used is of course due to the increase in the number of  
 381 sites on nanosorbent available for binding  $\text{Ag}^+$ . The lowest efficiency observed ( $87.0 \pm 0.3\%$ ) is better than the 85 %  
 382 extraction efficiency reported in the literature (Abdolmohammad-Zadeh and Javan 2015) where a greater amount of  
 383 a similar nanosorbent was used. This difference was attributed to the higher surface area of the nanosorbent in contact  
 384 with the aqueous solution in this study compared with that in the literature (Abdolmohammad-Zadeh and Javan 2015),  
 385 as they confined their nanosorbent in a syringe while in this study, the nanosorbent was dispersed in the mixed metal  
 386 aqueous solution.

387 The selectivity factor of  $\text{Ag}^+$  over the interfering ions ( $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$ ) was also found to be dependent on nanosorbent  
 388 dose. From Fig. 11b, the highest selectivity for  $\text{Ag}^+$  over the interfering ions ( $K_{\text{Ag}^+/\text{Cu}^{2+}} = 4119.9$  and  $K_{\text{Ag}^+/\text{Pb}^{2+}} =$   
 389  $2250.9$ ) was obtained with the highest nanosorbent dose used (80 mg), although the reasons for this remain unclear.  
 390 Therefore, in practical terms, the highest purity of  $\text{Ag}^+$  can be recovered from a mixed metal solution containing the  
 391 ion alongside  $\text{Cu}^{2+}$  and  $\text{Pb}^{2+}$  by using as much of the nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  as would be  
 392 commercially possible.



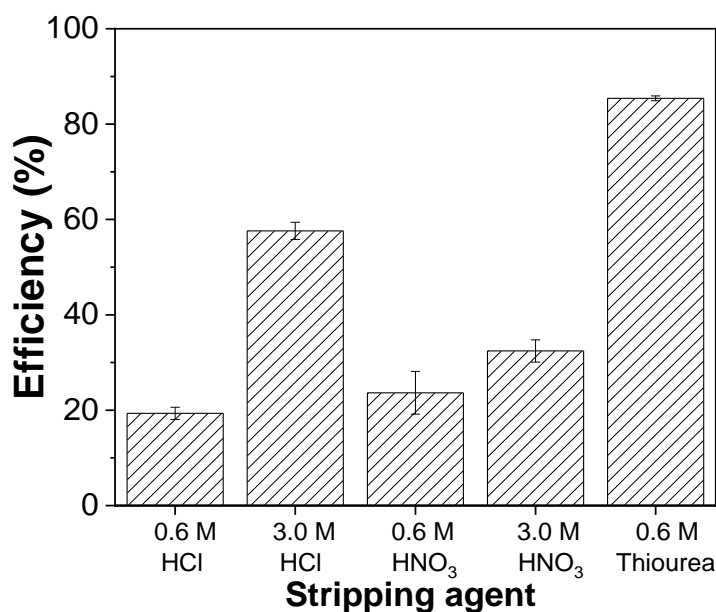
393  
 394 **Fig. 11** Effect of nanosorbent dose on (a) efficiency and (b) selectivity on  $\text{Ag}^+$  removal from simulated mixed metal  
 395 aqueous solution by the nanosorbent;  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  (conditions:  $[\text{M}^{n+}] = 2 \text{ mgL}^{-1}$ , volume = 10  
 396 mL, pH = 1.30, contact time = 45 min, temperature = RT).

397 Effect of type of stripping agents.

398 Recovery of the extracted metal ion is an important factor in studies on metal recycling. Furthermore, the choice of  
 399 an efficient metal recovery (or stripping) agent and establishment of optimum recovery conditions is important in  
 400 metal recycling studies. Therefore, the effect of the type of the stripping agent on the efficiency of  $\text{Ag}^+$  recovery was  
 401 investigated by contacting aqueous solutions of three different types of stripping agents – HCl (0.6 and 3 M),  $\text{HNO}_3$   
 402 (0.6 and 3 M) and thiourea (0.6 M) with  $\text{Ag}^+$  loaded nanosorbent -  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$ , noting that 3 M

403 thiourea could not be prepared due to the limited solubility in water. The choice of HCl and thiourea for this study  
404 was informed by previous work which reported excellent stripping efficiencies of  $\text{Ag}^+$  by these stripping agents  
405 (Abdolmohammad-Zadeh and Javan 2015; Vojoudi et al. 2017; Kazemi, Haji Shabani, and Dadfarnia 2015; Shimojo  
406 and Goto 2004).

407 Stripping of  $\text{Ag}^+$  from the  $\text{Ag}^+$ -impregnated nanosorbent was found to be dependent on the type of stripping agent.  
408 Predictably, the highest % SE ( $85.4 \pm 1.3$  %) was observed for 0.6 M thiourea and this was attributed to the preference  
409 of the soft *S*- donor atom in thiourea for the soft  $\text{Ag}^+$  acceptor (Fig. 12) (Pearson 1968). Presumably, a higher % SE  
410 could be observed at higher thiourea concentration. Interestingly, 3 M, HCl as found to strip more  $\text{Ag}^+$  than  $\text{HNO}_3$  of  
411 the same concentration (HCl =  $59.5 \pm 5.5$  % vs  $\text{HNO}_3$  =  $30.9 \pm 3.4$  %). This could be attributed to the formation of the  
412 anionic complex –  $\text{AgCl}_2^-$  at high  $\text{Cl}^-$  concentrations (Abdolmohammad-Zadeh and Javan 2015). The % SE observed  
413 for both acids at 0.6 M are identical considering their errors. Generally, an increase in the concentration of the stripping  
414 agent (excluding thiourea) resulted in an increase in % SE, but thiourea was the agent of choice.



415  
416 **Fig. 12** Relationship between stripping agent type and concentration in the stripping of  $\text{Ag}^+$  from impregnated  
417  $\text{Fe}_3\text{O}_4@\text{SiO}_2@[\text{MTESPIIm}]^+[\text{Cl}]^-$  nanosorbent (volume of acid = 5 mL, contact time = 1 h, temperature = RT).

#### 418 **Conclusion**

419 Surface functionalized magneto-responsive nanoparticles are a new class of materials which have received much  
420 interest in recent years for their potential utility as next generation agents for the recovery of metal(loid)s from the  
421 aqueous phase. Herein the ligand  $[\text{MTESPIIm}]^+[\text{Cl}]^-$  on  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  was prepared and characterized using FTIR, XPS,  
422 TGA, TEM and SQUID which confirmed that the nanomaterial exhibited a well constrained composition, physical  
423 structure and particle size distribution, in addition to a high saturation magnetism. SQUID analysis recorded a

424 magnetic saturation of  $50.30 \pm 0.10$  emu/g and thereby indicating that the nanomaterial is highly amenable for its  
425 transport and recovery from the aqueous phase using an externally applied magnetic field. Characterization using TGA  
426 recorded a surface coverage of the ionic liquid coating on the  $\text{Fe}_3\text{O}_4@/\text{SiO}_2$  of approximately 2 molecules/nm<sup>2</sup>. The  
427 [MTESPIIm]<sup>+</sup>[Cl]<sup>-</sup> on  $\text{Fe}_3\text{O}_4@/\text{SiO}_2$  was found to be highly efficient and selective for Ag<sup>+</sup> removal from solutions also  
428 containing Cu<sup>2+</sup> and Pb<sup>2+</sup>, with optimum efficiency and selectivity recorded at pH 3, with an exposure time of between  
429 0-15 minutes. Analysis of the Langmuir isotherm indicated a monolayer coverage of Ag<sup>+</sup> on the nanosorbent, with a  
430 saturation capacity of 23.69 mg/g. Thiourea (0.6 M) was the most effective Ag<sup>+</sup> stripping agent, which is attributed to  
431 the preference of the soft S- donor atom in thiourea for the soft Ag<sup>+</sup> acceptor. Overall, the results confirm that  
432 [MTESPIIm]<sup>+</sup>[Cl]<sup>-</sup> on  $\text{Fe}_3\text{O}_4@/\text{SiO}_2$  is a highly effective, versatile and selective agent for the removal of Ag from the  
433 aqueous phase and is therefore well suited for multiple future applications in both waste water treatment and mining  
434 sectors.

435

436 **Funding.** The authors would like to thank The UK Commonwealth Scholarship Commission for funding the research  
437 under grant number NGCS-2015-448.

438 **Compliance with ethical standards**

439 **Conflict of Interest.** The authors declare that they have no conflict of interest.

440

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